

Journal of Environmental Science and Health, Part B



Pesticides, Food Contaminants, and Agricultural Wastes

ISSN: 0360-1234 (Print) 1532-4109 (Online) Journal homepage: http://www.tandfonline.com/loi/lesb20

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To cite this article: Ellen T. Chang & Elizabeth Delzell (2016) Systematic review and meta-analysis of glyphosate exposure and risk of lymphohematopoietic cancers, Journal of Environmental Science and Health, Part B, 51:6, 402-434, DOI: 10.1080/03601234.2016.1142748

To link to this article: http://dx.doi.org/10.1080/03601234.2016.1142748

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Systematic review and meta-analysis of glyphosate exposure and risk of lymphohematopoietic cancers

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ABSTRACT

This systematic review and meta-analysis rigorously examines the relationship between glyphosate exposure and risk of lymphohematopoietic cancer (LHC) including NHL, Hodgkin lymphoma (HL), multiple myeloma (MM), and leukemia. Meta-relative risks (meta-RRs) were positive and marginally statistically significant for the association between any versus no use of glyphosate and risk of NHL (meta-RR = 1.3, 95% confidence interval (Cl) = 1.0–1.6, based on six studies) and MM (meta-RR = 1.4, 95% Cl = 1.0–1.9; four studies). Associations were statistically null for HL (meta-RR = 1.1, 95% Cl = 0.7–1.6; two studies), leukemia (meta-RR = 1.0, 95% Cl = 0.6–1.5; three studies), and NHL subtypes except B-cell lymphoma (two studies each). Bias and confounding may account for observed associations. Meta-analysis is constrained by few studies and a crude exposure metric, while the overall body of literature is methodologically limited and findings are not strong or consistent. Thus, a causal relationship has not been established between glyphosate exposure and risk of any type of LHC.

ARTICLE HISTORY

Received 28 November 2013

KEYWORDS

Glyphosate; non-Hodgkin lymphoma; Hodgkin lymphoma; multiple myeloma; leukemia; hematologic malignancies; herbicides; meta-analysis

Introduction

The broad-spectrum herbicide glyphosate (*N*-(phosphonomethyl)glycine), as a constituent of more than 750 products for agricultural, forestry, urban, and residential applications, is the most commonly used herbicide in the world. Therefore, understanding its potential human carcinogenicity has major implications for public health and risk assessment.

In 2014, the German Federal Institute for Risk Assessment (BfR), on behalf of the European Union, reviewed all toxicological studies of glyphosate in laboratory animals, as well as over 30 epidemiological studies in humans, and concluded that "the available data do not show carcinogenic or mutagenic properties of glyphosate" and "there is no validated or significant relationship between exposure to glyphosate and an increased risk of non-Hodgkin lymphoma or other types of cancer." ^[1,2] This conclusion was consistent with those previously reached by the United States Environmental Protection Agency (U.S. EPA) and the Joint Meeting on Pesticide Residues (JMPR), sponsored by the Food and Agriculture Organization of the United Nations and the World Health Organization (WHO), which concluded that glyphosate was unlikely to be carcinogenic to humans. ^[3–5]

By contrast, the International Agency for Research on Cancer (IARC) in 2015 classified glyphosate as "probably carcinogenic to humans" (Group 2A). In arriving at this classification, IARC characterized evidence of carcinogenicity in humans as "limited," based on the data available for non-Hodgkin lymphoma (NHL). [6] IARC considered the evidence of carcinogenicity in experimental animals as "sufficient." The latter determination was based on the occurrence of renal tubule

carcinoma, hemangiosarcoma, and pancreatic islet-cell adenoma in rodents, as well as mechanistic evidence.

To incorporate the IARC classification into the European Union review of glyphosate, BfR was commissioned by the German government and the European Food Safety Authority (EFSA) to review the IARC assessment.^[7] In its subsequent revised assessment report, BfR reached the conclusion that "no carcinogenic risk to humans is to be expected from glyphosate if it is used in the proper manner for the intended purpose." This assessment was supported by all European Union member states except one (Sweden) and by EFSA. ^[9] The WHO also has established an expert taskforce to re-evaluate the available data on glyphosate and report its findings to JMPR. ^[10]

In summarizing the epidemiological evidence, IARC stated that "case-control studies in the USA, Canada, and Sweden reported increased risks for NHL associated with exposure to glyphosate. The increased risk persisted in the studies that adjusted for exposure to other pesticides. The [Agricultural Health Study cohort did not show an excess of NHL. The Working Group noted that there were excesses reported for multiple myeloma in three studies; however, they did not weight this evidence as strongly as that of NHL because of the possibility that chance could not be excluded; none of the risk estimates were statistically significant nor were they adjusted for other pesticide exposures."^[6] A recent meta-analysis conducted by investigators from IARC^[11] found a statistically significant positive association between glyphosate use and NHL risk (meta-relative risk [RR] = 1.5, 95% confidence interval [CI] = 1.1-2.0), based on six studies. [12-17] The same metaanalysis also found a significant positive association between



glyphosate use and risk of B-cell NHL, based on two studies.[14,18]

Although Schinasi and Leon^[11] stated that in their metaanalysis, "[i]n an effort to use the most unbiased estimate, [they] extracted the most adjusted effect estimate," two or arguably three of the RR estimates that they selected for inclusion were not the most highly adjusted estimates reported by the original authors. [13-15] Instead, in a personal communication (11 August 2015), Dr. Schinasi indicated that other estimates were selected based on considerations of consistency of estimates across meta-analyses of other pesticides, secondary analyses, and statistical modeling approach.

Meta-analyses are not intended to identify, validate, or dispute causal relationships. Although they can be useful in providing a summary measure of association and identifying heterogeneity among research results, they can obscure important differences in methods and results among studies that can be more thoroughly evaluated in a detailed qualitative review. Schinasi and Leon[11] did not assess study quality and did not specifically address the potential impact of study limitations on the findings for glyphosate, nor did they discuss whether the apparent association between glyphosate and NHL risk is likely to be causal. On the other hand, Mink et al. [19] conducted a qualitative systematic review, without a meta-analysis, of epidemiologic studies of glyphosate and various cancers, including NHL. Taking into account potential sources of error, including selection bias, confounding, and especially exposure misclassification, the authors concluded that they "found no consistent pattern of positive associations indicating a causal relationship between total cancer (in adults or children) or any site-specific cancer and exposure to glyphosate."

Given the conflicting findings surrounding this issue, we conducted this systematic review and meta-analysis to examine more rigorously the relationship between exposure to glyphosate and risk of NHL, as well as major histopathological subtypes of NHL, in human epidemiologic studies. Because NHL is often considered alongside other lymphohematopoietic cancers (LHC), whose ever-changing classification systems now characterize some leukemias and multiple myeloma (MM) as NHL subtypes, [20] we also included Hodgkin lymphoma (HL), MM, and leukemia in this review. Despite the limitations of quantitative meta-analysis for observational epidemiology, [21,22] we conducted a meta-analysis largely to determine the impact of using RR estimates not used in the meta-analysis by Schinasi and Leon. [11] In addition, we conducted a qualitative evaluation of potential for error and bias. Thus, this article goes beyond previous work by examining all types of LHC, conducting a new meta-analysis, providing a detailed evaluation of study quality and potential for bias, and synthesizing the overall epidemiologic evidence for a causal association between glyphosate and LHC risk.

Methods

Literature search

Sources eligible for inclusion in the meta-analysis were original articles describing epidemiological studies that provided numeric point estimates of the RR (i.e., odds ratio, rate ratio, or prevalence ratio) of LHC, including NHL, HL, MM, leukemia, and any subtypes of these disease entities, associated with individual-level glyphosate exposure, along with corresponding interval estimates (e.g., 95% confidence intervals [CI]) or sufficient raw data to calculate RRs and CIs. Reviews, commentaries, letters to the editor without original data, and non-human studies were excluded, as were articles that did not report quantitative measures of association between glyphosate exposure (e.g., those assessing broadly defined categories of pesticides or herbicides) and risk of LHC (e.g., those assessing other cancers or all malignancies combined).

To identify all potentially relevant articles, we searched MEDLINE via PubMed (Supplementary methods), with additional targeted searches in Web of Science and Google Scholar, along with a review of the bibliographies of recent review articles. Based on a review of titles and abstracts to exclude articles without pertinent information, followed by a review of the full text of relevant articles, 19 articles (as well as one letter to the editor^[23] that contained additional results from a study described in another one of the included articles, [24] and one abstract^[25] that preceded a full-length article^[26]) were ultimately deemed eligible for inclusion (Appendix Fig. A1). Two authors independently reviewed and agreed upon the list of eligible articles.

Of the 19 articles reporting on the association between glyphosate and risk of specific forms of LHC, 12 pertained to NHL or its subtypes (including hairy-cell leukemia, which is a subtype of B-cell NHL), [12–18,24,27–30] 2 pertained to HL, [17,31] 6 pertained to MM, [12,17,26,32–34] and 3 pertained to leukemia. [12,35,36]

Evaluation of study characteristics and quality

From each eligible study, we extracted the following information: first author, publication year, study location, study design, study years, source population, number of subjects, proportion of proxy respondents, exposure assessment method, outcome assessment method, confounders adjusted, number of subjects in each exposure category, and RR estimates with CIs.

In addition to summarizing study characteristics, we qualitatively evaluated the methodological quality of each study in terms of its potential for selection bias, information bias/exposure misclassification, confounding, reporting bias, and other issues affecting validity. Potential for bias was evaluated based on subject identification strategy, participation rates, investigator blinding, assessment methods for exposures, outcomes, and potential confounders, statistical approach, reporting of results, and other considerations. [37-39]

Selection of data for meta-analysis

From each publication, we selected an RR point estimate for inclusion in the meta-analysis based on a set of rules specified a priori. First, if unadjusted and adjusted RRs were reported in a publication or across multiple publications from the same study population, the most fully adjusted RR was selected for inclusion. The most fully adjusted RR was defined as the RR estimate that took into consideration, by restriction or statistical adjustment, the most covariates that appeared to be confounders. The rationale for choosing the most fully adjusted RR was

| Authors | Year | RR | 95% CI | | 1- | | Relative weight (%) |
|----------------------|------|------|-------------|-----|----------------|-----|---------------------|
| De Roos et al. [13] | 2003 | 1.6 | 0.9-2.8 | - 1 | ├ ■─ | - 1 | 16.2 |
| De Roos et al. [12] | 2005 | 1.1 | 0.7-1.9 | - 1 | - ₱- | | 21.0 |
| Eriksson et al. [14] | 2008 | 1.51 | 0.77 - 2.94 | - 1 | +=- | | 11.6 |
| Hardell et al. [15] | 2002 | 1.85 | 0.55-6.20 | - 1 | - - | - | 3.6 |
| McDuffie et al. [16] | 2001 | 1.20 | 0.83 - 1.74 | - 1 | # | | 38.1 |
| Orsi et al. [17] | 2009 | 1.0 | 0.5 - 2.2 | - 1 | -+- | | 9.5 |
| Meta-RR | | 1.3 | 1.0-1.6 | - 1 | • | | |
| | | | | 0.1 | 1.0 | 10 | |

Figure 1. Forest plots of relative risk (RR) estimates and 95% confidence intervals (Cls) for the association between glyphosate exposure and risk of non-Hodgkin lymphoma. Meta-RRs were identical in random-effects and fixed-effects models.

based on the assumption that the adjusted covariates were found by the authors to act as confounders by altering the estimate of association (either directly or by acting as a surrogate for another, unmeasured confounder); however, some authors did not explain how confounders were selected, so this assumption may not hold for all studies. If an adjusted RR was not reported, the unadjusted (crude) RR was included as reported by the authors or as calculated from available raw data. Second, if multiple eligible publications were derived from the same study population, the RR from the most recent publication was selected for inclusion unless it was based on a subset of the overall eligible study population, in which case the RR based on the most complete study population was included. Third, subject to the first two rules, the RR for dichotomous exposure with the largest number of exposed cases was selected for inclusion in the meta-analysis. In a few instances where another RR from a given study nearly met these inclusion criteria but was superseded by a more fully adjusted, more recent, or more robust RR, the alternative RR was considered in secondary analyses.

RRs for multiple categories of exposure also were extracted to enable qualitative evaluation of exposure-response trends (based on the assumption, discussed later, that studies were able to distinguish among exposure levels). However, because no two studies used the same set of three or more categories to classify glyphosate exposure, these estimates could not be combined in meta-analysis.

Statistical approach

For associations with at least two independent RR estimates from different study populations, we estimated both fixedeffects and random-effects meta-RRs with 95% CIs. We used comparison of meta-RR estimates from fixed-effects and random-effects models as one approach to the evaluation of the impact of between-study heterogeneity on the meta-RRs. As a quantitative measure of between-study heterogeneity, we calculated I^2 , which represents the percentage of between-study variance in RRs that is attributable to study heterogeneity (as opposed to chance).^[40] We also tested for statistically significant between-study heterogeneity based on Cochran's Q statistic, [41] although this test has low power to detect modest heterogeneity across a limited number of studies.^[42]

In the absence of statistically significant heterogeneity, the presence of at least one statistically significant association, I^2 50%, and at least four contributing studies, we evaluated evidence of publication bias (i.e., non-random selection of studies

for publication, with a tendency toward submission and publication of studies that report larger, statistically significant associations^[43]) by using the linear regression approach of Egger et al., [44] which measures the degree of funnel plot asymmetry. We also estimated meta-RRs corrected for publication bias by imputing results for missing studies using the trim-and-fill procedure developed by Duval and Tweedie, [45] which iteratively trims asymmetric studies from the overbalanced side of a funnel plot to locate the unbiased effect, and then fills the plot by re-inserting the trimmed studies on the original side of the mean effect, along with their imputed counterparts on the opposite side. Again, we used these approaches with the understanding that they have limited power to detect publication bias based on few studies. [42]

The meta-analysis was conducted using Comprehensive Meta-Analysis Software (Biostat, Inc., Englewood, NJ, USA). All calculated meta-RRs and 95% CIs were confirmed using Episheet (www.krothman.org/episheet.xls).

Sensitivity analysis

To evaluate the robustness of results to various potential sources of heterogeneity, we planned a priori to conduct a sensitivity analysis with stratification of studies by study design (casecontrol vs. cohort), source of controls (population-based vs. hospital-based), gender (males only vs. males and females), geographic region (North America vs. Europe), and time period of cancer diagnosis (1980s, 1990s, or 2000s, with studies contributing to a given stratum if any part of the case diagnosis period was in a given decade).

Overall evaluation

To guide a qualitative assessment of the combined epidemiologic evidence for a causal relationship between glyphosate exposure and risk of LHC, we used Sir Austin Bradford Hill's "viewpoints" as a general framework. [46] Because this review is restricted to the epidemiologic literature, our consideration of the biological plausibility of the association and the coherence of the human, animal, and mechanistic evidence was limited.

Results

Study characteristics and overlap

Studies of NHL and subtypes

Twelve studies from seven independent study populations, including eleven case-control studies and one prospective cohort study, evaluated the relationship between glyphosate use and risk of NHL and/or its histopathological subtvpes. [12-18,24,27-30] Characteristics of these studies are summarized in Table 1. All of the studies considered glyphosate use in agricultural operations or settings, and most evaluated overall NHL as an outcome. The exceptions were Cocco et al., [18] which analyzed B-cell lymphoma and other NHL subtypes, but not overall NHL, and Nordstrom et al., [30] which included only hairy-cell leukemia. Eriksson et al. [14] presented results for B-cell lymphoma and other NHL subtypes, as well as for overall NHL, while Orsi et al.[17] included results for overall NHL and several specific NHL subtypes.

De Roos et al. [13] combined data from Cantor et al. [24] with data from two other studies that did not independently report associations between glyphosate use and NHL risk; [47,48] therefore, we did not further consider Cantor et al. [24] as a separate study. Lee et al. [29] was based on Cantor et al. [24] and Hoar Zahm et al., [48] but not Hoar et al., [47] and stratified results by asthma status (with no apparent interaction between glyphosate exposure and asthma); therefore, results from De Roos et al.^[13] took precedence in our analysis over those from Lee et al. [29] The study by Hardell et al. [15] pooled data from two other studies that reported on glyphosate use and NHL risk. [27,30] Consequently, the latter two studies were not considered further with respect to NHL, although Nordstrom et al. [30] was evaluated separately with respect to hairy-cell leukemia. Based on the same study population as McDuffie et al.[16] (except for four fewer cases excluded after pathology review), Hohenadel et al. [28] reported associations with use of glyphosate with or without malathion, but not glyphosate overall; therefore, the results from McDuffie et al. [16] were prioritized in our

The seven independent studies ranged markedly in size with respect to the number of NHL cases classified as exposed to glyphosate (based on reported use): Cocco et al., [18] 4 B-cell lymphoma cases exposed; Hardell et al., [15] 8 exposed; Orsi et al.,^[17] 12 exposed; Eriksson et al.,^[14] 29 exposed; De Roos et al., [13] 36 exposed; McDuffie et al., [16] 51 exposed; De Roos et al., [12] 71 exposed in the total eligible cohort. Four studies were based in Europe^[14,15,17,18] and three in North America[12,13,16] (Table 1). Four of the case-control studies were population-based, [13-16] one was hospital-based, [17] and one included a mixture of population-based and hospital-based cases and controls.[18] Four studies were restricted to males, [13,15-17] while the rest included males and females. Two studies conducted at least some case ascertainment during the 1980s, [13,15] five during the 1990s, [12,14-16,18] and four during the 2000s^[12,14,17,18] (categories are overlapping). For reference, glyphosate entered the U.S. and European commercial markets in 1974.^[49]

Studies of HL

Two case-control studies estimated the OR between glyphosate use and risk of HL.[17,31] Characteristics of these studies are summarized in Table 1. The study by Karunanayake et al.^[31] used the same methods and source population as McDuffie et al., [16] but focused on HL rather than NHL.

As described in the section on NHL studies, Orsi et al. [17] was a hospital-based case-control study set in Europe (France), restricted to males, with case ascertainment in the 2000s, participation rates > 90%, and no proxy respondents. This study classified six HL cases as exposed to glyphosate. Karunanayake et al. [31] was a population-based case-control study set in North America (Canada), restricted to males, with case ascertainment in the 1990s, participation rates of 68% for cases and 48% for controls, and an unspecified proportion of proxy respondents. In this study, 38 HL cases were classified as glyphosate-exposed.

Studies of MM

Six studies from four independent study populations, including four case-control studies and two prospective cohort studies, evaluated the association between glyphosate use and risk of MM. [12,17,26,32-34] These studies are described in Table 1. A cross-sectional analysis within a subset of the Agricultural Health Study Cohort examined the association between glyphosate use and risk of monoclonal gammopathy of unknown significance (MGUS), an MM precursor; [50] this study was not included in the present review.

The studies by De Roos et al. [12] and Sorahan [26] were based on virtually identical datasets from the Agricultural Health Study cohort (except that the dataset used by Sorahan was stripped of data on race, state of residence, and applicator type due to privacy concerns; these differences should not have affected the results substantively). Because the Sorahan^[26] study included all eligible cohort members, whereas the De Roos et al. [12] study was based on a restricted subset of the cohort with complete data, [51] the Sorahan [26] results were prioritized in our analysis of MM. Brown et al. [32] employed the same methods and source population as Cantor et al., [24] which was included in the pooled analysis of NHL by De Roos et al. [13] Pahwa et al. [34] and Kachuri et al. [33] conducted overlapping analyses in the same Canadian source population as McDuffie et al., [16] Hohenadel et al., [28] and Karunanayake et al. [31] Pahwa et al. [34] included more controls in their analysis, but these controls were excluded from Kachuri et al. [33] because they were younger than any enrolled MM cases (≤29 years) and thus did not contribute meaningfully to the analysis. Kachuri et al. [33] also controlled for more confounders, and therefore was prioritized in our analysis.

With respect to glyphosate use, the four independent studies of MM included, respectively, 5 exposed cases, [17] 11 exposed cases, [32] 24 exposed cases, [26] and 32 exposed cases. [33] All but one study, which was based in France, [17] were conducted in North America, and all except one^[26] were restricted to males. One of the two case-control studies was population-based^[32] and the other was hospital-based. [17] Case ascertainment took place during the early 1980s in one study, [32] at least partly during the 1990s in two studies, [26,33] and at least partly during the 2000s in two studies.[17,26]

Studies of leukemia

Two case-control studies and one prospective cohort study investigated the relationship between glyphosate use and risk of leukemia. [12,35,36] Key characteristics of these studies are provided in Table 1. The study by Brown et al. [35] used the same methods

Table 1. Design characteristics of studies of glyphosate exposure and risk of lymphohematopoietic cancer (LHC), including non-Hodgkin lymphoma (NHL), NHL subtypes, Hodgkin lymphoma (HL), multiple myeloma (MM), and leukemia.

| leukemia. | | | | | | | | | | |
|---------------------------------|------|---|--|--|---|---|--|--|--|--|
| Authors | Year | Outcomes studied | Study location | Study design | Study years | Source population | Subject identification | Subject participation | Subjects (n) | Proxy respondents |
| Brown et al. ^[35] | 1990 | Leukemia (including myelod/splasias) | United States (lowa and Minnesota) | Population-based case-control | 1980–1983 | White men aged ≥ 30 years in lowa and Minnesota, excluding Minnesotis, St. Paul, Duluth, and Rochester | Cases: lowa Tumor Registry and special surveillance of Minnesota hospital and pathology laboratory records. Controls: random-digit claiming if aged < 65 years, Medicare files if aged < 65 years, state death | Cases: 86% Controls: 77% random digit dialing, 79% Medicare, 77% proxise for deceased proxise for deceased supplemental interview: 93% cases, 96% controls | Cases: 578 Controls: 1,245 Supplemental interview: 86 cases, 203 controls | Cases: 238 (41%) Controls: 425 (34%) Supplemental interview, 63 (73%) cases, 57 (28%) controls |
| Brown et al. ^[32] | 1993 | MM | United States (lowa) | Population-based case-control | 1981–1984 | White men aged \geq 30 years in lowa | centificate files if deceased Cases: lowa Health Registry Controls: random-digit dialing if aged < 65 years, Medicare files if aged > 65 years, state death | Cases: 84% Controls: 78% overall | Cases: 173 Controls: 650 | Cases: 72 (42%) Controls: 198 (30%) |
| Cantor et al. ^[24] | 1992 | NHL | United States (lowa and Minnesota) | Population-based case-control | 1980–1983 | White men aged ≥ 30 years in lowa and Minnesota, excluding Minnesolis, St. Paul, Duluth, and Rochester | certificates if deceased cases: lowa State Health Registry and special surveillance of Minnesota hospital and pathology laboratory records pathology laboratory records opents: and open deceased to see a s | Cases: 89% Controls: 77% random-digit dialing, 79% Medicare, 77% proxies for deceased | Cases: 622 Controls: 1245 | Cases: 184 (30%) Controls: 425 (34%) |
| Cocco et al. ^[18] | 2013 | B≺ell NH.L | Europe (Czech Republic, France, Germany, Ireland, Italy, and Spain) | Population- and hospital-based case-control | 1998–2004 | Persons aged > 17 years in Germany and Italy general populations, and in referral areas of participating hospitals in Czech Republic, France, Ireland, and Spain | Gases. NR Controls: random sampling of population registers in Germany and Italy; recruitment from hospital departments for infectious and parasitic (17.6%), mental and nervous (14.6%), mental and nervous (14.6%), endocrine and metabolic (7.1%), endocrine and metabolic (4.1%), respiratory (3.9%), and several other conditions (33.2%), excluding cancer, in Czech Republic, France, Ireland, and | Cases: 88% overall; 90% Czech Republic, 91% France, 87% Germany, 90% Ireland, 93% Italy, 82% Spain Controls: 69% overall, 81% hospital-based, 52% population- based; 60% Cere Republic, 74% France, 44% Germany, 75% Ireland, 66% Italy, 96% Spain | Cases: 2348 Controls: 2462 | None |
| De Roos et al. [13] | 2003 | NHL | United States (Nebraska, Jowa, Minnesota, and Kansas) | Population-based case-control (pooled analysis of 3 studies) | 1979–1986 | White men aged \geq 21 years in one of the 66 counties of eastem Nebraska; white men aged \geq 30 years in lowa and Minnesota, excluding Minneapolis, St. Paul, Duluth, and Rochester; white men aged \geq 21 years in Kansas | Spania Group and area hospitals; lowa Group and area hospitals; lowa State Health Registry; special surveillance of Minnesota hospital and pathology laboratory records; University of Kansas Cancer Data Service registry Controls: random-digit dialing if aged < 65 years, state death if aged > 65 years, state death certificate files if deceased | Cases: 91% Nebraska (93% living, 89% deceased); 89% kanas Minnescus; 96% Kanasa Controls: 85% Nebraska; 77% random-digit dialing, 79% Medicare, 77% deceased (proxies) lowa and Minnescha; 93% Kanasa Analysis restricted to subjects who lived or worked on a farm before 18 years of age (% NR); analysis of multiple pesticides restricted to subjects with non-missing | Cases: 650 (in analyses of multiple pesticides) Controls: 1933 (in analyses of multiple pesticides) | Cases: 201 (30,9%) (in analyses of multiple pesticides) Controls: 767 (39,7%) (in analyses of multiple pesticides) |
| De Roos et al. ^[1,2] | 2005 | LHC, NHL, MM, leukemia | United States (lowa and North Carolina) | Prospective cohort | 1993–1997 through 2001 Median = 6.7 years | Private and commercial pesticide applicators in lowa and North Carolina who were licensed to apply restricted-use pesticides | Pesticide applicators identified when seeking a state-issued restricted-use pesticide license; invited to complete the invited to complete the enrollment questionnaire at the licensing facility | data (75% cases, 75% controls) 298 subjects (0.5%) lost to follow- up or with no person-time contributed > 80% of eligible pesticide applicators enrolled in study by completing on-site questionnaire 44% of applicators completed | Eligible cohort: 36,509–49,211 in analyses adjusted for demographics and lifestyle 30,613–40,719 in analyses additionally adjusted for other pesticides | None e |
| Eriksson et al. ^[14] | 2008 | NH, B-cell NH, SLL/CLL, FL grades I-III, DLBCL, other specified B-cell NH, unspecified B-cell NHI, T-cell NHL, unspecified NHL | Europe (Sweden) | Population-based case-control | 1999–2002 | Adults aged 18–74 years in 4 of 7 health service regions in Sweden associated with university hospitals in Lund, Linköping, Örebro, and Umeå | Cases: contact with treating physicians and pathologists Controls: national population registry | take-home questionnaire Gases: 81% Controls: 65% (92% of initially enrolled controls with 71% participation) | Cases: 995 Controls: 1016 | None |

| Cases: 177 (44%) Controls: NR (~44%; matched to cases) | Cases: ~35% (NR) Controls: ~29% (NR) | Gases; 110 (21%) Controls: 220 (15%) | Cases: 103 (30%) Controls: 202 (15%) | Gases: NR Controls: 220 (15%) | Q. | Cases: 266 (31%) Controls: 779 (33%) |
|--|---|--|--|--|---|---|
| Con | Con | Case | Case | Сом | None | Case |
| Cases: 404 Controls: 741 | Cases: 515 Controls: 1141 | Cases: 513 Controls: 1506 | Gases: 342 Controls: 1357 | Cases: 316 Controls: 1506 | Cases: 180 Controls: 756 | Cases: 872 Controls: 2336 |
| Cases: 91% (91% living, 92% deceased) Controls: 84% (83% living, 85% deceased) | Cases: 91% Controls: 84% | Cases; 67% Controls: 48% Based on postal codes, respondents were not more or less likely than non-respondents to live in a rural area. | Cases: 58% Controls: 48% Based on postal codes, respondents were not more or less likely than non-respondents to live in a rural area. | Cases: 68% Controls: 48% Based on postal codes, respondents were not more or less likely than non-respondents to live in a rural area. | Cases: 100% Controls: 100% | Cases: 91% Nebraska, 89% lowa and Minnesota Controls: 85% Nebraska, 78% lowa and Minnesota |
| Cases: regional cancer registries Controls: national population registry if living, national registry for causes of death if deceased | Cases: regional cancer registries for NHL, national cancer registry for hairy-cell leukemia Controls: national population registry, national registry for causes of death if deceased | Cases: hospital records in Quebec, cancer registries in all other provinces controls: provincial health insurance records in Alberta, Saskatrowan, Maritoba, and Quebec; computerized telephone listings in Ontario; | voter lists in British Columbia clases: hospital records in Quebec, cancer registries in all other provinces controls: provincial health insurance records in Alberta, Saskatchewan, Maintoba, and Quebec; computerized telephone listings in Ontario; | voter lists in British Columbia cases: hospita records in Quebec, cancer registries in all other provinces controls: provincial health insurance records in Alberta, Saskatchevan, Manitoba, and Quebec; computerized telephone listings in Ontario; | voter lists in British Columbia Cases: hospital records for acute infection or inflammation (33%), trauma (22%), acute abdominal emergencies such as appendicitis (27%), or various other diagnoses with elective admission, such as cataract, henia repair, or cosmetic surgery (17%), excluding head trauma with loss of consciousness or cancer, controls at Dhomburi Hospital (a nearby pprivate hospital) matched to 21 cases admitted to private wards. | for wealthy patients Cases: Nebrask by tymphoma Study Group and area hospitals; lowa State Health Registry; special surveillance of Minnesota hospital and pathology laboratory records Controls: random/dgit daling if aged < 65 years. Medicare files if aged ≥ 65 years, state death certificate files if deceased |
| Men aged ≥ 25 years in the four northernmost counties of Sweden and three counties in mid- | Sweden Men aged ≥ 25 years in the four northermost counties of Sweden and three counties in mid- Sweden (for NHL) or in the entire country of Sweden (for hairy-cell | ieukemia) Men aged ≥ 19 years in Alberta, British Columbia, Manitoba, Ontario, Quebec, and Saskatchewan | Men aged ≥ 19 years (≥ 30 years in analysis) in Alberta British Columbia, Mantioba, Ontario, Quebec, and Saskarchewan | Men aged ≥ 19 years in Alberta, Birtish Columbia, Manitoba, Ontario, Quebec, and Saskatchewan | Patients aged ≥ 18 years residing in Bangkok proper and suburbs of Nonthaburri, Abrompathom, Patumthani, Samutprakam, and Samutprakam, and Samutprakam, and to Sirraj Hospital or Dhonburi Hospital | White men and women aged > 21 years in one of 45 counties in eastern Nebraska, white men aged > 30 years in lowa and Minnesota, excluding Minneapolis, 5r. Paul, Duluth, and Rochester |
| 1987–1990 | 1987–1990 | 1991–1994 | 1991–1994 | 1991–1994 | 1997–2003 | 1980–1986 |
| Population-based case-control | Population-based case-control | Population-based case-control | Population-based case-control | Population-based case-control | Hospital-based case-control | Population-based case-control (pooled analysis of 2 studies) |
| Europe (Sweden) | Europe (Sweden) | Canada (Alberta, British Columbia, Manitoba, Ontario, Quebec, and Saskatchewan) | Canada (Alberta, British Columbia, Manitoba, Ontario, Quebec, and Saskatchewan) | Canada (Alberta, British Columbia, Manitoba, Ontario, Quebec, and Saskatchewan) | Bangkok, Thailand | United States (Nebraska, lowa, and Minnesota) |
| JHN | NHL including hairy-cell leukemla | NHL | MM | 보 | Leukemia | 보 |
| 1999 | 2002 | 2011 | 2013 | 2012 | 5009 | 2004 |
| Hardell and Eriksson ^[27] | Hardell et al. ¹¹⁵¹ | Hohenadel et al. ^[28] | Kachuri et al. ^[33] | Karunanayake et al. ¹³¹ l | Kaufman et al. ^[36] | Lee et al. ^[29] |

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| Ϋ́ | Year Outcomes studied | Study location | Study design | Study years | Source population | Subject identification | Subject participation | Subjects (n) | Proxy respondents |
| McDuffie et al. ^[1:0] 20 | 2001 NHL | Canada (Alberta, British Columbia, Manitoba, Ontario, Quebec, and Saskatchewan) | Population-based case-control | 1991–1994 | Men aged ≥ 19 years in Alberta, British Columbia, Manitoba, Ontario, Quebec, and Saskatchewan | Cases: hospital records in Quebec, cancer registries in all other provinces Controls; provincial health insurance records in Alberta, Saskarchevan, Manitoba, and Quebec; computerized telebrone listing in Ontraio; vereleprone listing in Ontraio; vereleprone listing finish Columbia | Cases: 67% Controls: 48% Based on postal codes, respondents were not more or less likely than non-respondents to live in a rural area. | Cases: 517 Controls: 1506 | Cases: ~21% (NR) Controls: 220 (15%) |
| Nordström 19 et al. ^[30] | 1998 Hairy-cell leukemia | Europe (Sweden) | Population-based case-control | 1987–1992 (1993 for | Men living in Sweden | Cases: national cancer registry Controls: national population | Cases: 91% Controls: 83% | Cases: 111 Controls: 400 | Cases: 4 (4%) Controls: 5 (1%) |
| | 2009 LHC, NHL, DLBCL, FL, IPS, CLL, hairy-cell leukemia, HL, MM | Europe (France) | Hospital-based case- control | 2000–2004 | Men aged 20–75 years living in the catchment areas of the main hospitals in Brest, Gaen, Nantes, Lille, Toulouse, and Bordeaux, with no history of immunosuppression or taking immunosuppressant drugs | Cases: hospital records Controls: hospital records for controls: hospital records for orthopedic or rheumatological conditions (89.3%), gastrointestinal or genitourinary ract diseases (4.8%), ardiovascular diseases (1.1%), skin and subcutaneous tissue disease (1.8%), and infections (3.0%), excluding patients admitted for cancer or a disease directly related to occupation, smoking, or alcohol abuse | Cases: 95.7% Controls: 91.2% | Cases: 491 LHC, 244 NHL, 104 LPS, 87 HL, 56 MM Controls: 456 | None |
| 34 | 2012 MM | Canada (Alberta, British Columbia, Maniroba, Ontario, Quebec, and Saskarchewan) | Population-based case-control | 1991–1994 | Men aged ≥ 19 years in Alberta, British Columbia, Manitoba, Ontario, Quebec, and Saskatchewan | Gases: hospital records in Quebec, cancer registries in all other provinces Controls: provincial health insurance records in Alberta, Saskatchewan, Manitoba, and Quebec, computerized telephone listings in Ontario; voter lists in British Columbia | Cases: 58% Controls: 48% Based on postal codes, respondents were not more or less likely than non-respondents to live in a rural area. | Cases: 342 Controls: 1506 | Cases: 103 (30%) Controls: 220 (15%) |
| Sorahan ⁽²⁶⁾ 20 | 2015 MM | United States (lowa and North Carolina) | Prospective cohort | 1993–1997 through 2001 Median = 6.7 years | Private and commercial pesticle applicators in lowa and North Carolina who were licensed to apply restricted-use pesticides | Pesticide applicators identified when seeking a state-issued restricted-use pesticide licens; invited to complete the enrollment questionnaire at the licensing facility | 298 subjects (0.5%) lost to follow- up or with no person-time contributed > 80% of eligible pesticide applicators emrolled in study by completing on-site questionnaire 43% of applicators completed take-home questionnaire | Eligible cohort (1): 54,315 educling subjects with cancer before enrollment, loss to follow-up, missing age at enrollment, or missing glyphosate use 49,211 also excluding missing education, smoking, or alcohol 40,719 excluding missing other pesticides Eligible cohort (2): 53,656 excluding subjects with cancer before enrollment, loss to follow-up, missing age at enrollment, missing age at enrollment, missing age at enrollment, or glyphosate use or missing cumulative exposure days of glyphosate use Eligible cohort (3): 55,934 excluding aubjects with cancer before enrollment, and and a public cohort (3): 55,934 excluding aubjects with cancer before enrollment, loss to follow-up, or missing age at enrollment | None |

| columns). |
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| (additional |
| Continued |
| Table 1. |

| Authors | Year | Exposure assessment | Outcome assessment | Investigator blinding | Confounders considered or adjusted | Funding source | Overlap |
|---------------------------------|------|--|---|---|---|---|---|
| Brown et al. ^[35] | 1990 | In-person structured interview, including detailed farming and pesticide use history. For each pesticide, evaluated ever use, first and last year of use, and personal applying/mixing/handling handling. In 1987, supplemental telephone interview to evaluate usual number of days of pesticide use per year among lows subjects who had reported and of roadin parezides. | Diagnostic confirmation by regional pathologists, special review of myelodysplasias by one pathologist co-author | 2 | Adjusted: vital status, age, state, ever used tobacco dally, first-degree family history of LHC, non-farming job related to leukemia risk in this study, exposure to substances (benzene, naphtha, hair dyes) related to leukemia risk in this study | Partial support from National Institute of Environmental Health Sciences | Brown et al. ^[32] , Cantor et al. ^[34] , De Roos et al. ^[13] , Lee et al. ^[28] |
| Brown et al. ^[32] | 1993 | In-person structured interview, including detailed farming and pesticide use history for each pesticide, evaluated ever use, first and last year of use, person (use, person) and use of concerne annimons. | Diagnostic confirmation by an expert pathologist | 8 | Adjusted: vital status, age Considered: smoking, education, other factors found not to be confounders of agricultural risk factors | Partial support from National Institute of Environmental Health Sciences | Brown et al. ¹³⁵ ; Cantor et al. ¹³⁴ ; De Roos et al. ¹³³ ; Lee et al. ²³ |
| Cantor et al. ^[24] | 1992 | In-person structured interview, including detailed farming and pesticide use history of all subjects who had worked on a farm for > 6 months since age 18 years. For each pesticide, evaluated ever use, first and last year of use, method of application, personal applying/mixing/handling, and use of protective continuous. | Diagnostic confirmation and morphological dassification by panel of 4 experienced regional pathologists | °Z | Adjusted: vital status, state, age, cigarette smoking status, first-degree family history of LHC, non-faming job related to NHL risk in this study, exposure to hair dyes, exposure to other substances associated with NHL risk in this study Considered; pesticides belonging to other chemical families | Partial support from National Institute of Environmental Health Sciences | Brown et al. ^[35] ; Brown et al. ^[32] ; De Roos et al. ^[13] ; Lee et al. ^[29] |
| Cocco et al. ^[18] | 2013 | In-person structured interview, including detailed farming and pesticide use history for all subjects who reported having worked in agriculture. For each agricultural job, reported tasks, crops, size of cultivated area, pest treated, pesticides used, crop treatment procedures, use of personal procedures, use of personal procedures, the entry after treatment, and frequency of freatment in days not vesser. | Histologically or oytologically confirmed cases with central review of sides of ~20% by an international team of pathologists | ^o Z | Adjusted: age, gender, education, study center | European Commission, 5th and 6th Framework Programmes; Spanish Ministry of Health; German Federal Office for Radiation Protection; La Fondation de France; Italian Ministry for Education, University and Research; Italian Association for Cancer Research | None |
| De Roos et al. ^[13] | 2003 | Telephone interview in Nebraska and Kansas, Inperson structured interview in lowa and Minnesota manages. In Menaska: Question about use of any pesticide, followed by prompting for specific selected pesticides, including years of use and average days per year lowa and Minnesota: Direct question about a selected use of specific pesticides, including first and last years of use a selected use of specific pesticides, including first and last years of use for specific pesticides, followed by questions on duration of use and days per year for groups of pesticides but not including interview. | Nebraska: Pathology review with histological confirmation and dassification including immunologic phenotyping lowa and Minnesota. Diagnostic confirmation and morphological classification by panel of 4 experienced regional pathologists Kansas. Diagnostic confirmation and dassification by panel of 3 pathologists | Yes in Nebraska; no in lowa, Minnesota, and Kansas and Kansas | Adjusted: age, study site, other individual pesticides with \geq 20 users in full study Considered: first-degree family history of LHC, education, smoking | NR; assume National Cancer Institute | Brown et al. ^[32] ; Brown et al. ^[32] , Cantor et al. ^[42] ; Lee et al. ^[22] (also Hoar et al. ^[47] , Hoar Zahm et al. ^[46]) |
| De Roos et al, ^{1,2} | 2005 | Self-administered written questionnaire (with validation study) evaluating detailed use of 22 pesticides for private applicators, 28 pesticides for commercial applicators, 8 pesticides for commercial applicators (ever/mever use, frequency, duration, and intensity of use, decade of first use), and even/mever use for additional pesticides up to total of 50, with general information no pesticide application methods, personal protective equipment, pesticide mixing, and equipment repair. Additional self-administered take-home questionnaire with further questions on corrunational expressions and exceptional expressions. | Linkage to state cancer registry files, state death registries, and National Death Index | None | Adjusted: age at enrollment, education, cigarette smeking pack-years, alcohol consumption in past year, first-degree family history of cancer, state of residence. Considered (adjusted for MM only): 5 pesticides for which canualitive evasoure-days were most which canualitive evasoure-days were most highly associated with those for glyphosate (i.e., 24-dichlorophenoxyacetic acid, alachlor, atrazine, metobachor, trifluralins), 5 pesticides for which ever/never use was most highly associated with that for glyphosate (i.e., benomyl, maneb, paraquat, carbaryl, diazinon) | National Cancer Institute, National Institute of Environmental Health Sciences, Environmental Protection Agency, and National institute for Occupational Safety and Health | Sorahan (261 |
| Eriksson et al. ^[14] | 2008 | Self-administered mailed questionnaire with additional telephone interview for missing or unclear answers; evaluated occupationel exposure to individual pesticides, including number of years, number of days per year, and approximate length of exposure per day | Diagnostic pathological specimens examined and dassified by 1 of 5 Swedish expert lymphoma reference pathologists, if not already initially reviewed by one of them; pane I review if dassification differed from original report | Yes | Adjusted: age, sex, and year of diagnosis or enrollment; other associated agents (4-chloro-2- methyl phenoxyacetic acid, 2.4- dichlorophenoxyacetic acid and/or 2.4,5- trichlorophenoxyacetic acid, mercurial seed dressing, arsenic, creosote, tar) for NHL only | Swedish Council for Working Life and Social Research; Carcer and Allergy Fund; Key Fund; Örebro University Hospital Cancer Fund | None |

| I able 1. (Continued) | | | | | | | |
|--------------------------------------|------|---|--|---|--|--|--|
| Authors | Year | Exposure assessment | Outcome assessment | Investigator blinding | Confounders considered or adjusted | Funding source | Overlap |
| Hardell and Eriksson ⁽²²⁾ | 1999 | Self-administered mailed questionnaire with supplemental telephone interview for unclear answers, assessed use of pesticides within different occupations, were consert if not handling the sprayer, brand names of pesticides, years of exposure, and cumulative days of exposure. Exposure excluded 1 year prior to diagnosis or index | Histopathological diagnosis of NHL reported to regional cancer registries, confirmed by review of pathology reports | Yes | Adjusted: age, county, vital status, year of death if deceased, use of phenoxyacetic acids | Swedish Work Environment Fund, Swedish Medical Research Council, Örebro County Council Research Committee, Örebro Medical Center Research Foundation | Hardell et al. ^[15] |
| Hardell et al. ^[15] | 2002 | year Sef-administered mailed questionnaire with supplemental telephone interview for unclear answers; assessed years and total number of days of occupational exposure to various agents and names of agents. | Histologically verified NHL: confirmation of hairy-cell leukemia NR | Yes | Adjusted: study, study area, vital status, other associated pesticides (4-chloro-2-methyl phenoxyacetic acid, 2,4-dichlorophenoxyacetic acid + 2,4-5-trichlorophenoxyacetic acid, other herbicides) | Swedish Cancer Research Fund, Swedish Medical Research Council, Örebro County Council Research Committee, Örebro Medical Centre Research Foundation | Hardell and Eriksson ^[27] Nordström et al. ^[30] |
| Hohenadel et al. ^[28] | 2011 | period or ≥ 1 year Telephone interview for detailed information on pesticide use in subjects who reported in a self- administered mail questionaire that they had ≥ 10 hours of pesticide use during their lifetime, plus 1596 random sample of subjects with < 10 hours Pesticide interview (with validation study) included a pre-mailed list of specific pesticides (chemical and trade names) with number of days used and number of hours per day at home or work for | Diagnostic confirmation based on information, including pathology reports, from cancer registries and hospitals; pathological and hospitals; pathological anterial reviewed and classified by a reference pathologist; subjects with unavailable pathological material retained in study | 2 | Adjusted: age, province, use of a proxy respondent Considered: diesel exhaust, ultraviolet radiation, farm animals, chemicals such as benzene, first-degree family history of cancer | Health Canada, British Columbia Health Research Foundation, Centre for Agricultural Medicine at University of Saskatchewan | Kachuri et al. ^[33] . Karunanayake et al. ^[31] . McDuffe et al. ^[16] . Pahwa et al. ^[34] |
| Kachuri et al ^{133]} | 2013 | Telephone interview for detailed information on pesticide use in subjects who reported in a self-administrated mail questionaire that they had ≥ 10 hours of pesticide use during their lifetime, plus 15% andom sample of subjects with < 10 hours Pesticide interview (with validation study) included a pre-mailed list of specific pesticides (chemical and trade names) with number of days used and number of hours per day at home or work for each nearities. | Diagnostic confirmation based on information, including pathology reports, from cancer registries and hospitals, pathological material reviewed and dassified by a reference pathologist (including pathology and tumor tissue sides for 125 (37%) of 34 cases); subjects with unavailable pathological material retained in pathological material retained in | 2 | Adjusted: age, province, use of a proxy respondent, smoking status, personal history of rheumatoid arthifis, allergies, measles, shingles, or cancer, family history of cancer | Occupational Cancer Research Centre; Cancer Care Ontario; Ontario Workplace Safety and Insurance Board; Canadian Cancer Society, Ontario Division, Mitacs- Accelerate Graduate Research Internship Program | Hohenadel et al. ^[28] , Karunanayake et al. ^[31] , McDuffie et al. ^[16] , Pahwa et al. ^[34] |
| Karunanayake et al. ^[31] | 2012 | Telephone interview for detailed information on pesticide use in subjects who reported in a self-administed mall questionnaire that they administed mall questionnaire that they had ≥ 10 hoursyvan of cumulative exposure to any combination of herbicides, insecticides, fungacides, fungacides, fungants, and algicides. Pesticide interview collected information on exposure to individual pesticides, place of pesticide use, year of first use, first year on market, number of years of use, and days per year of use [Note differences from related studies] | Initial diagnosis based on information from cancer registries and hospitals, pathology and turn of tissue sildee for 155 of 316 cases reviewed by a reference pathologist who confirmed HL in 150/155 cases, plus 7 cases originally classified as NHL; subjects with unavailable pathological material retained in pathological material retained in | 2 | Adjusted: age, province, personal history of measles, acre, hay fever, or shingles, first-degree family history of cancer | NR, assume same as in related studies | Hohenadel et al. ^[28] . Kachuri et al. ^[33] . McDuffe et al. ^[16] . Pahwa et al. ^[58] |
| Kaufman et al. ^[36] | 2009 | Interview with nurse to assess occupational and non-occupational exposure to pesticides and other potential risk factors | Histologically confirmed leukemia diagnosed within 6 months before current hospital attendance or admission | ° | Considered: age, sex, income, use of cellular telephones, benzene and other solvent exposure, occupational and non-occupational pesticide exposure, pesticides used near home, working with power lines, living near power lines, exposure to X-rays, exposure to carlain types of plantromannair fields use of hir dwas | Thailand Research Fund and Commission on Higher Education | None |
| Lee et al. ^[29] | 2004 | Telephone interview in Nebraska; in-person structured interview in lowa and Minnesota Questions included personal handling of groups of pesticides and individual pesticides used on crops or animals, with years of first and last use | Nebraska: Pathology review with histological confirmation and classification including immunologic phenotyping lowa and Minnesota. Diagnostic confirmation and morphological classification by panel of 4 experienced regional | Yes in Nebraska; no in lowa and Minnesota | Adjusted: age, state vital status Considered: gender, smoking, first-degree family history of LHC, ever having a job correlated with risk of LHC (e.g., painting or welding), use of protective equipment | NR, assume National Cancer Institute | Brown et al. ^[32] . Brown et al. ^[32] . Cantor et al. ^[34] . De Roos et al. ^[134] . (also Hoar Zahm et al. ^{[48)}) |

| Hohenadel et al [^{28]} , Karburi et al [^{13]} ; Karunanayake et al [^{31]} , Pahwa et al [^{34]} | Hardell et al. ^[15] | None | Hohenadel et al. ^[28] ; Kachuri et al. ^[33] ; Karunanayake et al. ^[31] ; McDuffie et al. ^[16] |
|--|---|---|--|
| Health Canada, British Columbia Health Research Foundation, Centre for Agricultural Medicine at University of Saskatchewan | Swedish Work Environment Fund, Örebro County Council Research Committee, Örebro Medical Centre Research Foundation. | Association pour la Recherche contre le Cancer, Fondation de France, AFSSET, Faberge employees (donation) | Occupational Cancer Research Centre; Cancer Care Ontario, Ontario Workplace Safety and Insurance Board; Canadian Cancer Society |
| Adjusted: age, province, personal history of measles, mumps, cancer, or allegy desensitization shots, first-degree family history of cancer. Considered: pesticide exposure, smoking history | Adjusted: age Considered: exposure to animals, herbicides, insecticides, fungicides, impregnating agents, organic solvents, exhausts, or ultraviolet light | Adjusted: age, study center, socioeconomic category Considered: all combinations of pesticide families associated with the LHC subtype considered with a p-value = 0.10, rural/urban status, type of housing, educational level, history of monoru-cleosis, history of influenza immunization, family history of cancer, skin characteristics, smoking status, and alcohol drinking status | Adjusted: age, province, personal history of measles, mumps, allergies, arthritis, or shingles, first-degree family history of cancer |
| 2 | Yes | Š | 2 |
| Diagnostic confirmation from cancer registries and hospitals; pathological material reviewed and classified by a reference pathologics; subjects with unavailable pathological material retained in study | Reported to national cancer registry; further confirmation not described | All diagnoses cytologically or histologically confirmed and reviewed by a panel of pathologists and hematologists | Diagnostic confirmation based on information, including pathology reports, from cancer registries and hospitals; pathological material reviewed and classified by a reference pathologist (including pathology and tumor tissue sides for 125 [37%] of 342 cases); subjects with unavailable pathological material retained in study |
| Telephone interview for detailed information on pesticide use in subjects who reported in a self-administered mail questionmaire that they had \geq 10 hours of pesticide use during their lifetime, plus 15% random sample of subjects with < 10 hours (total = 179 case, 456 controls with telephone interview) with validation study) included a pre-mailed list of specific pesticides (chemical and trade names) with number of days used and number of hours per day at home or work for pesticides. | Self-administered mailed questionnaire with supplemental telephone interview for unclear or missing answers; assessed total number of days of occupational exposure defined as 2 I working day with influction period of > 1 wear | Self-administered written questionnaire with lifetime accupational history, followed by inperson structured interview evaluating non-occupational instory, followed by inperson structured interview evaluating non-occupational instory followed by inperson structured interview evaluating non-worked as a farmer or gardener for subjects who had worked as a farmer or gardener for self-months, period of occupation and area, farmer's status at each farm, crops and animal husbandry with mean sizes, all pesticides used on each crop during a given period, whether subject had personally prepared, mixed, or sprayed the pesticide, chemical used, annual number and duration of applications, and use of pesticides in farm buildings for animals, grain, hay or straw, or to clear lares and yards. All questionnaires reviewed by an occupational hygienits and an agronomist, networms from 95 (56.8%) of 158 subjects who completed the agricultural questionnaire, not completed by 355 (20.8%) who refused (n = 15), died/were in poor health (n = 10), or could not be contacted (n = 15); all chancing to edder using ad hoc system and dascefind as defining or poscible approach | Telephone interview for detailed information on pesticide use in subjects who reported in a self-administered mail questionnaire that they had ≥ 10 h of pesticide use during their lifetime, plus 15% random sample of subjects with < 10 h Pesticide interview (with validation study) included a pre-mailed list of specific pesticides (chemical and rade names) with number of days used and number of hours per day at home or work for each pesticide |
| 2001 | 1998 | 5000 | 2012 |
| . McDuffie et al. ^[16] | Nordström et al. ^[30] | Orsi et al. ^[17] | Pahwa et al ^[34] |

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|-------------------------|------|--|--|-----------------------|---|-----------------------|--------------------------------|
| Authors | Year | Exposure assessment | Outcome assessment | Investigator blinding | Confounders considered or adjusted | Funding source | Overlap |
| Sorahan ^[26] | 2015 | Self-administered written questionnaire (with validation study) evaluating detailed use of 22 pesticides for private applicators, 28 pesticides for commercial applicators (seve/never use, frequency, duration, and intensity of use, decade of first use), and ever/never use for additional pesticides up to tolar of 50, with general information on pesticide application methods, personal protective equipment, pesticide mixing, and equipment repair. Additional self-administered take-home questionnaire with further questions on occupational exposures and lifestyle factors Missing data classified into 'not known/missing' caregoty, with unknown use of 2.4—dichlorophenoxyacetic acid classified with no use and unknown education classified with no education beyond high school due to lack of MM cases in unknown education and use to accept the self-additional exposures and due to lack of MM cases in unknown education and use of actions of the self-additional exposures and due to lack of MM cases in unknown education and use of actions of the self-additional exposures and easily action desirile due to lack of MM cases in unknown education and easile due to lack of MM cases in unknown education and entered and entered and unknown education and entered and entered and entered and entered and entered entered and entered ente | Linkage to state cancer registry files, state death registries, and National Death Index | None | Fully adjusted: age, gender, smoking pack-years, alcohol tuse in year before enrollment, first-degree family history of cancer, education, use of 2,4-dichlorophenoxyacate acid, alculor, atrazine, metolachtor, or trifluralin, ever use of benomy, maneb, paraquat, carbaryl, or diazinon intermediate adjusted; age, gender, smoking, alcohol, family history of cancer, education Adjusted in full cohort: age, gender, family history of cancer, education | Monsanto Europe SA/NV | De Roos et al. ^{17.2} |

G: confidence interval; CLL: chronic lymphocytic leukemia; DLBCL: diffuse large B-cell lymphoma; FL: follicular lymphoma; HL: Hodgkin lymphoma; LHC: lymphohematopoietic cancer; LPS: lymphoproliferative syndrome; MM: multiple myeloma; NHL: non-Hodgkin lymphoma; NR: not reported; OR: odds ratio; SLL: small lymphocytic lymphoma.



and source population as Brown et al., [32] which was described in the section on MM, and Cantor et al., [24] which was included as part of De Roos et al. [13] in a pooled analysis of NHL.

As described earlier, De Roos et al., [12] the only prospective cohort study included, was based in North America (Iowa and North Carolina), enrolled both males and females, ascertained cancer incidence in the 1990s and 2000s, and had a 99.5% follow-up rate through 2001. In the total eligible cohort, 43 leukemia cases occurred among glyphosate users. Brown et al. [35] was a population-based case-control study set in North America (Iowa and Minnesota), restricted to white males, with cases identified in 1980-1983, participation rates of 86% for cases and 77-79% for controls, and proxy respondent rates of 41% for cases and 34% for controls. Fifteen leukemia cases in this study were classified as having used glyphosate. The other casecontrol study of leukemia, by Kaufman et al., [36] was a hospitalbased study set in Asia (Thailand), with males and females, case ascertainment in the 1990s and 2000s, participation rates of 100%, and no proxy respondents for cases or controls.

Meta-analysis

NHL

All relevant RRs and 95% CIs for the association between reported glyphosate use and risk of overall NHL, including those not used in the meta-analysis, such as estimates within subgroups, minimally adjusted estimates, and estimates of exposure-response patterns, are provided in Table 2. The estimates selected from each independent study population for inclusion in the meta-analysis, according to the rules specified in the methods section, are provided in Table 3.

As shown in Table 3 and Fig. 1, the combined meta-RR for overall NHL in association with any use of glyphosate, based on six studies, [12-17] was 1.3 (95% CI = 1.0-1.6). The results were identical in the random-effects and fixedeffects models, suggesting limited between-study heterogeneity in the association. Little heterogeneity also was indicated by the I^2 value of 0.0% and the highly nonsignificant P-value of 0.84 for Cochran's Q. Given the lack of heterogeneity and at least one statistically significant association, we tested for publication bias using Egger's linear regression approach to evaluating funnel plot asymmetry, and found no significant asymmetry (one-tailed Pvalue = 0.20). Using Duval and Tweedie's trim-and-fill approach to adjust for publication bias, the imputed meta-RR for both the random-effects and fixed-effects models was 1.2 (95% CI = 1.0-1.6).

In secondary analyses, we replaced the RR estimated by De Roos et al.^[13] using a hierarchical (i.e., multistage) regression model with the RR estimated using a more traditional logistic regression model (Table 3). (The hierarchical regression RR was selected for the primary analysis because, as stated by the authors, hierarchical regression models can yield "increased precision and accuracy for the ensemble of estimates" when modeling multiple pesticides simultaneously, and the more conservative prior assumptions specified in these models "seemed appropriate in a largely exploratory analysis of multiple exposures for which there is little prior knowledge about how

pesticide exposures interact in relation to the risk of NHL.") Using the logistic regression RR did not appreciably affect the results of the meta-analysis (meta-RR = 1.3, 95% CI = 1.0-1.6; identical for random-effects and fixed-effects models).

In another secondary analysis, we replaced the RR reported by McDuffie et al. [16] with the results reported by Hohenadel et al. [28] in the same study population (minus four previously misclassified NHL cases) (Table 3). Because Hohenadel et al. [28] reported two estimates for glyphosate use—one in the absence of malathion use and one in the presence of malathion use—we combined these two estimates into a single estimate (RR = 1.40, 95% CI = 0.62-3.15) using random-effects meta-analysis. Using this alternative estimate also did not appreciably affect the meta-RR (1.3, 95% CI = 1.0-1.7; identical for randomeffects and fixed-effects models). Finally, using both the logistic regression RR instead of the hierarchical regression RR from De Roos et al. [13] and the combined RR from Hohenadel et al.^[28] instead of the RR from McDuffie et al.^[16] slightly but non-significantly increased the meta-RR to 1.4 (95% CI = 1.0-1.8; identical for random-effects and fixed-effects models) (Table 3).

As noted earlier, in their meta-analysis of the association between glyphosate use and NHL risk, Schinasi and Leon^[11] included RR estimates from Eriksson et al.^[14] and Hardell et al. [15] that were not the most highly adjusted estimates reported by the authors (shown in Table 2 as univariate odds ratios). They also used the logistic regression estimate from De Roos et al.[13] that arguably was not as highly adjusted as the hierarchical regression estimate. When we included these estimates in the meta-analysis, along with the same estimates from De Roos et al., [13] McDuffie et al., [16] and Orsi et al. [17] as included in our main meta-analysis, we obtained the same results as reported by Schinasi and Leon: [11] random-effects meta-RR = 1.5, 95% CI = 1.1-2.0 (I^2 = 32.7%, $p_{heterogeneity}$ = 0.19). The fixed-effects meta-RR based on these estimates (not reported by Schinasi and Leon^[11]) was 1.4 (95% CI = 1.1-1.8).

NHL subtypes

All reported RRs and 95% CIs for the association between glyphosate use and risk of various NHL subtypes are shown in Table 2. The estimates included in meta-analyses, which were conducted for B-cell lymphoma, diffuse large B-cell lymphoma, chronic lymphocytic leukemia/small lymphocytic lymphoma, follicular lymphoma, and hairy-cell leukemia (i.e., all NHL subtypes for which at least two estimates from independent studies were available), are shown in Table 3. Too few studies of any given NHL subtype were conducted to justify testing for publication bias.

The meta-RR for the association between any use of glyphosate and risk of B-cell lymphoma, based on two studies, [14,18] was 2.0 (95% CI = 1.1-3.6) according to both the random-effects and the fixed-effects model ($I^2 = 0.0\%$, p_{he}terogeneity = 0.58) (Table 3). These results are the same as reported by Schinasi and Leon.[11] The four B-cell lymphoma cases who were classified by Cocco et al. [18] as having used glyphosate consisted of one patient with diffuse large B-cell lymphoma, one with chronic lymphocytic

+ Table 2. Estimated associations between glyphosate exposure and risk of lymphohematopoietic cancer (LHC), including non-Hodgkin lymphoma (NHL), NHL subtypes, Hodgkin lymphoma (HL), multiple myeloma (MM), and leukemia.

| leukemia. | | | | |
|--------------------------------------|------|--|--|--|
| Authors | Year | Exposure groups and number of subjects | Relative risk | D %56 |
| Brown et al. ^[35] | 1990 | Non-farmers: 243 cases, 547 controls | Leukemia OR = 0.9 | Leukemia 95% CI = 0.5-1.6 |
| Brown et al. ^[32] | 1993 | ever mixed, nandled, or applied glypnosate: 15 cases, 49 controls Non-farmers: 62 cases, 272 controls | MMOR = 1.7 | MM 95% CI = 0.8-3.6 |
| Cantor et al. ^[24] | 1992 | Ever mixed, handled, or applied glyphosate: 11 cases, 40 controls Non-farmers: 226 cases, 547 controls | Among those who did not use protective equipment, MM OR = 1.9 NHL OR = 1.1 $$ | Among those who did not use protective equipment, MM 95% CI = NR NHL 95% CI = 0.7–1.9 |
| Cocco et al. ^[18] | 2013 | Ever handled, mixed, or applied glyphosate. 26 cases, 49 controls Unexposed to any pesticides: NR cases, 2262 controls Occupationally exposed to glyphosate: 4 cases (1 DLBCL, 1 CLL, 1 MM 1 increoffied Bacall NH1? 2 controls | B-cell NHL OR = 3.1 | B-cell NHL 95% CI = 0.6-17.1 |
| De Roos et al. [13] | 2003 | Unexposed to glyphosate: 614 cases, 1892 controls | Hierarchical regression NHL OR = 1.6 | Hierarchical regression NHL 95% CI = 0.9–2.8 |
| De Roos et al. ^[12] | 2005 | Exposed to glypnosate: 36 cases, 61 controls Never used glyphosate: 47 LHC, 21 NHL, 8 MM, 14 leukemia; | Logistic regression NHL OK = 2.1 Fully adjusted LHC RR = 1.1 | Logistic regression NHL 95% CI = 1.1 –4.0 Fully adjusted LHC 95% CI = 0.8–1.6 |
| | | 13,280 cohort members Ever used glyphosate 143 LHC 71 NHL 24 MM 43 leukemia | Age-adjusted LHC RR = 1.1 Fully adjusted NHI R8 = 1.1 | Age-adjusted LHC 95% CI = 0.8–1.5 Fully adjusted NHI 95% CI = 0.7–1.9 |
| | | 41,035 cohort members | Age-adjusted NHL RR = 1.2 | Age-adjusted NHL 95% CI = 0.7-1.9 |
| | | | Fully adjusted MM RR = 2.6 (2.6 in lowa, 2.7 in North Carolina) Age-adjusted MM RR = 1.1 | Fully adjusted MM 95% CI = 0.7–9.4 Age-adjusted MM 95% CI = 0.5–2.4 |
| | | | Fully adjusted leukemia RR = 1.0 | Fully adjusted leukemia 95% $CI = 0.5 - 1.9$ |
| | | 1_20 alvahasata avaosiira dave: 48 HC 29 NHI 8 MM 9 | Age-adjusted leukemia RR $= 1.1$ Cumulativa exposure dave tertiles 2 and 3 vs 1 | Age-adjusted leukemia 95% CI = 0.6–2.0 Cumulativa avaneura dave tartilae 2 and 3 ve 1 |
| | | leukemia | LHC RRs = 1.2 , 1.2 ; p-trend = 0.69 | LHC 95% CIs = 0.8–1.8, 0.8–1.8 |
| | | 21–56 glyphosate exposure days: 38 LHC, 15 NHL, 5 MM, 14 | NHL RRs = 0.7, 0.9; p-trend = 0.73 MM BRs = 1.1 10: n-frand = 0.27 | NHL 95% CIs = 0.4-1.4, 0.5-1.6 MM 95%, CIs = 0.4-3.5, 0.6-6.3 |
| | | 57–2,678 glyphosate exposure days: 36 LHC, 17 NHL, 6 MM, 9 | Leukemia RRs = 1.9, 1.0; p-trend = 0.61 | Leukemia 95% Cls = 0.8-4.5, 0.4-2.9 |
| | | leukemia | > 108 vs. > 0–9 exposure days, NHL RR = 0.9 | > 108 vs. > 0-9 exposure days, NHL 95% CI = 0.4-2.1 |
| | | 0.1–79.5 intensity-weighted glyphosate exposure days: 38 LHC, 24 NHL. 5 MM. 7 leukemia | Intensity-weighted exposure days, tertiles 2 and 3 vs. 1 LHC RRs $= 1.0.10$; p-trend $= 0.90$ | Intensity-weighted exposure days, tertiles 2 and 3 vs. 1 LHC 95% CIs $= 0.6-1.5, 0.7-1.6$ |
| | | 79.6–337.1 intensity-weighted glyphosate exposure days: 40 LHC, | NHL RRs = 0.6, 0.8; p-trend = 0.99 | NHL 95% CIs = 0.3-1.1, 0.5-1.4 |
| | | 15 NHL, 6 MM, 17 leukemia 337 2–18 241 intencity-weighted glunbocate exposure days: 43 | MM RRs = 1.2, 2.1; p-trend = 0.17 nukemia RRs = 1.9 | MM 95% CIs = 0.4-3.8, 0.6-7.0 outbandia 95% CIs = 0.8-4.7, 0.2-2, 1 |
| | | LHC, 22 NHL, 8 MM, 8 leukemia | Intensity tertile 3 vs. 1 | Intensity tertile 3 vs. 1 |
| | | | MM RR = 0.6 | MM 95% CI = 0.2–1.8 |
| | | | Cumulative exposure days, tertiles 1, 2, and 5 vs. never MM RRs = 2.3, 2.6, 44; p-trend = 0.09 | Cumulative exposure days, terules 1, 2, and 3 vs. never MM 95% Cls = 0.6-8.9, 0.6-11.5, 1.0-20.2 |
| | | | Cumulative exposure days, quartile 4 vs. never | Cumulative exposure days, quartile 4 vs. never |
| Fribecon of al [14] | auuc | No nacticida avanceura. ND | MM RR = 6.6; p-trend = 0.01 NHI OB any glyphocate multiposite = 1.51 | MM 95% CI = 1.4–30.6 NIHI 95% CI any alvabacata multivariata = 0.77-2.94 |
| Elikssoli et di. | 7000 | Glyphosate exposure for ≥ 1 full working day, ≥ 1 calendar year | NHL OR, any glyphosate, intuitivariate $=$ 1.5 I NHL OR, any glyphosate, univariate $=$ 2.02 | NHL 95% CI, any glyphiosate, illutivariate = 0.7 / 2.54 NHL 95% CI, any glyphosate, univariate = 1.10–3.71 |
| | | prior to year of diagnosis or enrollment: 29 NHL cases, 18 | NHL OR, glyphosate 1 to \leq 10 days = 1.69 | NHL 95% CI, glyphosate 1 to \leq 10 days = 0.70–4.07 |
| | | controls (NHL subtypes NR) Glynbocate expecting for 1 to $<$ 10 days: 12 | NHL OR, glyphosate > 10 days = 2.36 NHI OR any glyphosate latency 1–10 years — 1.11 | NHL 95% CI, glyphosate > 10 days = 1.04–5.37 NHL 95% CI any glymbosate latenty 1–10 waare = 0.24–5.08 |
| | | Offices, 9 controls | NHL OR, any glyphosate, latency > 10 years $= 2.26$ | NHL 95% CI, any glyphosate, latency > 10 years $= 0.27$ - 500 NHL 95% CI, any glyphosate, latency > 10 years $= 1.16$ - 4.40 |
| | | Glyphosate exposure for > 10 days: 17 NHL cases, 9 controls | B-cell NHL OR, any glyphosate = 1.87 | B-cell NHL 95% Cl, any qlyphosate = 0.998–3.51 |
| | | | SLL/CLL OR, any glyphosate $= 3.35$ | SLL/CLL 95% Cl, any glyphosate $= 1.42-7.89$ |
| | | | FL grades I–III OR, any glyphosate = 1.89 Dl RCI OR any glyphosate = 1.22 | FL grades I–III 95% CI, any glyphosate = 0.62–5.79 DI RCI 95% CI any glynhosate = 0.44–3.35 |
| | | | Other specified B-cell NHL OR, any glyphosate = 1.63 | Other specified B-cell NHL 95% CI, any glyphosate = $0.53-4.96$ |
| | | | Unspecified B-cell NHL OR, any glyphosate = 1.47 | Unspecified B-cell NHL 95% CI, any glyphosate = 0.33–6.61 |
| | | | ו-ניפון ואחר סיה, מווץ פוץ פון פון פון פון בענים בענים Unspecified NHL OR, any alvohosate = 5.63 | I-cell NHL 95% Ct, ally gryphosate = 0.51=10.4 Unspecified NHL 95% CI, any divphosate = 1.44–22.0 |
| Hardell and Eriksson ^[27] | 1999 | No pesticide exposure Glyphosate exposure > 1 wear prior to diagnosis or control index | NHL OR adjusted for phenoxyacetic acids = 5.8 NHL OR unadjusted for phenoxyacetic acids = 2.3 | NHL 95% CI adjusted for phenoxyacetic acids = 0.6–54 NHI 95% CI unadjusted for phenoxyacetic acids = 0.4–13 |
| | | year: 4 cases, 3 controls | | or to a second constant to the second constan |
| Hardell et al. ^[15] | 2002 | No pesticide exposure: NR Glyphosate exposure for ≥ 1 working day, ≥ 1 year prior to | Multivariate NHL OR = 1.85 Univariate NHL OR = 3.04 | Multivariate NHL 95% Cl = 0.55–6.20 Univariate NHL 95% Cl = 1.08–8.52 |
| [38] | | diagnosis or control index date: 8 cases, 8 controls | | |
| Honenadel et al | 7011 | Use of netther glyphosate nor malathon: 4.2. cases, 130 L controls Use of glyphosate only; 19 cases, 78 controls Use of malathion only: 41 cases 72 controls | NHL OK, glypnosate only = U32. NHL OR, malatrion only = U32. NH OR otwhosate and malatrion = 210 | NHL 95% (I, glyphosate only = 0.54~1.55 NHL 95% (I, malathion only = 1.29 ~2.93 NHI 95% (I nlynhosta and malathion = 131-337 |
| | | Use of glyphosate and malathion: 31 cases, 55 controls | Interaction contrast ratio = 0.23 , P-interaction = 0.69 | |

| MM 95% CI, ever glyphosate = 0.76 – 1.87 MM 95% CI, ever glyphosate, no proxies = 0.66 – 1.86 MM 95% CI, glyphosate > 0 to ≤ 2 days per year = 0.39 – 1.32 MM 95% CI, glyphosate > 0 to ≤ 2 days per year, no proxies = 0.35 – 1.40 MM 95% CI, glyphosate > 2 days per year = 0.98 – 4.70 MM 95% CI, glyphosate > 2 days per year, no proxies = 0.95 – 4.70 | Fully adjusted HL 95% CI = 0.62–1.56 Minimally adjusted (age, province) HL 95% CI = 0.74–1.76 Crude leukemia 95% CI = 0.15–13.56 | NHL 95% CI, non-farmers, asthmatics = 0.3-1.4 NHL 95% CI, glyphosate, non-asthmatics = 0.98–2.1 NHL 95% CI, glyphosate, asthmatics = 0.4-3.3 | Fully adjusted NHL 95% CI, ever glyphosate = $0.83-1.74$ Minimally adjusted (age, province) NHL 95% CI, ever glyphosate = $0.87-1.80$ Minimally adjusted (age, province) NHL 95% CI, glyphosate > 0 to ≤ 2 days per year = $0.63-1.57$ Minimally adjusted NHL 95% CI, glyphosate > 2 days per year = $1.20-3.73$ | Hairy-cell leukemia 95% ∟ = 0.8−1.2 | LHC 95% CI = 0.6–2.1 NHL 95% CI = 0.6–2.2 DUBCL 95% CI = 0.3–2.7 FL 95% CI = 0.4–5.2 LPS 95% CI = 0.1–1.8 Hair-cell endkemia 95% CI = 0.3–9.3 HI 95% CI = 0.6–5.0 | MM 95% CI = 0.8-7.3 MM 95% CI = 0.77-1.93 | Age- and sex-adjusted MM 95% CI, cohort of 54,315 = 0.52-2.94 Age- and sex-adjusted MM 95% CI, cohort of 54,315 = 0.82-2.41 Age-adjusted MM 95% CI, cohort of 54,315 = 0.82-2.41 Age-adjusted MM 95% CI, cohort of 49,211 = 0.66-5.53 Intermediate adjusted MM 95% CI, cohort of 49,211 = 0.71-6.04 Age-adjusted MM 95% CI, cohort of 40,719 = 0.62-7.48 Fully adjusted MM 95% CI, cohort of 40,719 = 0.62-7.48 Fully adjusted MM 95% CI, cohort of 40,719 = 0.62-4.86 Cumulative exposure days, tertiles 1, 2, and 3 vs. never Fully adjusted MM 95% CI, e. 0.44-2.80, 0.56-4.95, 0.42-3.58 Age- and sex-adjusted MM 95% CI = 0.42-2.70, 0.50-3.38, 0.33-3.11 Intensity-weighted exposure days, tertiles 1, 2, and 3 vs. never Fully adjusted MM 95% CI = 0.33-3.00, 0.45-3.50, 0.67-3.67 MM 95% CI, ever aljusted MM 95% CI = 0.31-2.62, 0.45-3.20, 0.57-3.67 MM 95% CI, ever aljusted MM 95% CI = 0.31-2.62, 0.42-3.00, 0.57-3.67 MM 95% CI, ever aljusted AMM 95% CI = 0.31-2.65 Cumulative exposure days, tertiles 1, 2, 3, and unknown vs. never MM 95% CI = 0.44-2.83, 0.54-3.88, 0.04-3.38, 0.04-3.56 MM 95% CI = 0.44-2.83, 0.54-3.88, 0.04-3.38, 0.04-3.98 MM 95% CI = 0.44-2.83, 0.04-3.19, 0.62-4.05, 0.22-4.92 |
|--|--|---|--|---|---|---|--|
| MM OR, ever glyphosate = 1.19 MM OR, ever glyphosate to proxies = 1.11 MM OR, glyphosate > 0 to \leq 2 days per year = 0.72 MM OR, glyphosate > 0 to \leq 2 days per year, no proxies = 0.70 MM OR, glyphosate > 2 days per year = 2.04 MM OR, glyphosate > 2 days per year = 2.04 MM OR, glyphosate > 2 days per year, no proxies = 2.11 | Fully adjusted HL $OR=0.99$ Minimally adjusted (age, province) HL $OR=1.14$ Crude leukemia $OR=1.40$ | NHL OR, non-farmers, asthmatics = 0.6 NHL OR, glyphosate, non-asthmatics = 1.4 NHL OR, glyphosate, asthmatics = 1.2 | Fully adjusted NHL OR, ever glyphosate = 1.20 whimmally adjusted (age, province) NHL OR, ever glyphosate = 1.26 Minimally adjusted (NHL OR, glyphosate > 0 to ≤ 2 days per year = 1.00 Minimally adjusted NHL OR, glyphosate > 2 days per year = 2.12 Minimally adjusted NHL OR, glyphosate > 2 days per year = 2.12 | Halfy-cell leukemia UK = 3.1 | LHC OR = 1.2 DIRL OR = 1.0 DIBCL OR = 1.0 F. OR = 1.4 DFS OR = 0.6 CL OR = 0.4 Harry-cell leukenia OR = 1.8 H. OR = 1.7 H. OR = 1.7 H. OR = 1.7 | WM OK = 2.4 MM OR = 1.22 | Fully adjusted MM RR, cohort of 54,315 = 1.24 Age - and sex-adjusted MM RR, cohort of 54,315 = 1.12 Age - and sex adjusted MM RR, cohort of 49,211 = 1.08 Age adjusted MM RR, cohort of 49,211 = 1.91 Intermediate adjusted MM RR, cohort of 49,211 = 1.207 Age-adjusted MM RR, cohort of 40,719 = 2.21 Fully adjusted MM RR, cohort of 40,719 = 2.22 Cumulative exposure days, tertiles 1, 2, and 3 vs. never Fully adjusted MM RRs = 1.13, 1.50, 1.23; p-trend = 0.48 using scores, o means Intermediate adjusted MM RRs = 1.06, 1.24, 1.08, p-trend = 0.50 using scores or means Intermediate adjusted MM RRs = 1.06, 1.24, 1.08, p-trend = 0.20 using scores on means Intermediate adjusted MM RRs = 0.91, 1.12, 1.44; p-trend = 0.22 using scores, 0.18 using means Intermediate adjusted MM RRs = 0.91, 1.12, 1.44; p-trend = 0.39 using scores, 0.33 using means MM RR, ever adjusted MM RRs = 0.91, 1.12, 1.43; p-trend = 0.39 using scores, 0.33 using means MM RR, unknown glyphosate = 1.77 Cumulative exposure days, tertiles 1, 2, 3, and unknown vs. never MM RR = 0.95, 1.19, 1.58, 1.04; p-trend = 0.30 using scores or means excluding unknown Internsity-weighted exposure days, tertiles 1, 2, 3, and unknown vs. never MM RR = 0.95, 1.19, 1.58, 1.04; p-trend = 0.30 using scores, 0.26 using means |
| Never used glyphosate: 310 cases, 1236 controls (216 cases, 1047 controls without proxy) Ever used glyphosate: 32 cases, 121 controls (23 cases, 108 controls without proxy) Used glyphosate for > 0 to < 2 days per year. 15 cases, 88 controls (11 cases, 78 controls without proxy) Used glyphosate for > 2 days per year. 12 cases, 29 controls (10 cases, 26 controls without proxy) | Never used glyphosate: 278 cases, 1373 controls Ever used glyphosate: 38 cases, 133 controls No glyphosate use: 179 cases, 753 controls Glyphosate: 1 case, 3 controls | Non-farmers, non-asthmatics: 259 cases, 684 controls Non-farmers, asthmatics: 9 cases, 37 controls Exposed to glyphosate, non-asthmatics: 53 cases, 91 controls Exposed to glyphosate, asthmatics: 6 cases, 12 controls | Never used glyphosate: 466 cases, 1373 controls between used glyphosate: 10 cases, 1500 controls Glyphosate use for > 0 to ≤ 2 days per year Glyphosate use for > 2 days per year | No glypnosate exposure: 10, cases, 359 controls Glyphosate exposure for 2 1 working day, 2 1 year prior to diagnosis or control index date: 4 cases, 5 controls | Never exposed to glyphosate: 464 LHC, 232 NHL, 102 DIBCL, 47 FL, 100 LPS, 75 CLL, 25 hairy-cell leukemia 81 HL, 51 MM, 432 controls. Ever exposed to glyphosate: 27 LHC, 12 NHL, 5 DLBCL, 3 FL, 4 LPS, 2 CLL, 2 hairy-cell leukemia, 6 HL, 5 MM, 24 controls | Never used glyphosate: 310 cases, 1373 controls | Ever used gyphosate: 8 cases, 13,280 cohord members (of 54,315); 4 cases, 11,881 cohord members (of 49,211); 3 cases, 9809 cohord members (of 40,719) Ever used gyphosate: 24 cases, 41,035 cohord members (of 54,315); 22 cases, 37,330 cohord members (of 49,211); 19 cases, 30,910 cohord members (of 49,211); 19 cases, 30,910 cohord members (of 49,211); 1-26 gyphosate exposure days: 10 cases 57-26/8 gyphosate exposure days: 6 cases 57-26/8 gyphosate exposure days: 6 cases 0.1-795 intensity-weighted glyphosate exposure days: 6 cases 336,-31, intensity-weighted glyphosate exposure days: 8 cases 337,2-18,241 intensity-weighted glyphosate exposure days: 10 cases 337,2-18,241 intensity-weighted glyphosate exposure days: 10 cases Unknown glyphosate use: 2 cases |
| 2013 | 2012 | 2004 | 2001 | 866 | 2009 | 2012 | 2015 |
| . Kachuri et al ^[83] | Karunanayake et al. [31] Kaufman et al. [36] | Lee et al. ^[29] | McDuffle et al. ^[16] | Nordstrom et al. | Orsi et al. ¹⁷⁷ | Pahwa et al. ^[34] | Sorahan ²⁸⁶ |

G: confidence interval; CLL: chronic lymphocytic leukemia; DLBCL: diffuse large B-cell lymphoma; FL: follicular lymphoma; HL: Hodgkin lymphoma; LHC: lymphohema; LPC: lymphopema; LPS: lymphopproliferative syndrome; MM: multiple myeloma; NR: not reported; OR: odds ratio; RR: relative risk; SLL: small lymphocytic lymphoma.



Table 3. Selected estimates included in meta-analyses and calculated meta-analysis relative risks (meta-RRs) of the association between glyphosate exposure and risk of (LHC), including non-Hodgkin lymphoma (NHL), NHL subtypes, Hodgkin lymphoma (HL), multiple myeloma (MM), and leukemia.

| Study # | Authors | Year | Outcome | Number of exposed subjects | RR | 95% CI | | |
|----------------------------|--|------------------------------|---|---|--|---|------------------------------|--------------------------------------|
| 1 | De Roos et al.[13] | 2003 | Non-Hodgkin lymphoma | 36 cases, 61 controls | a. 1.6 (hierarchical regression) b. 2.1 (logistic regression) | a. 0.9–2.8 (hierarchical regression) b. 1.1–4.0 (logistic regression) | | |
| 2 | De Roos et al.[12] | 2005 | Non-Hodgkin lymphoma | 71 cases* | 1.1 | 0.7–1.9 | | |
| | Eriksson et al.[14] | 2008 | Non-Hodgkin lymphoma | 29 cases, 18 controls | 1.51 | 0.77–2.94 | | |
| | Hardell et al.[15] | 2002 | Non-Hodgkin lymphoma | 8 cases, 8 controls | 1.85 | 0.55-6.20 | | |
| | Hohenadel et al.[28] | 2011 | Non-Hodgkin lymphoma | 50 cases, 133 controls | 1.40 (random effects meta-RR) | 0.62-3.15 | | |
| | Honenader et al. | 2011 | Non-Hougkin lymphoma | 30 cases, 133 controls | 1.40 (landom enects meta-ini) | (random effects meta-CI) | | |
| | McDuffie et al.[16] | 2001 | Non-Hodgkin lymphoma | 51 cases, 133 controls | 1.2 | 0.83–1.74 | | |
| , | Orsi et al. [17] | 2001 | Non-Hodgkin lymphoma | | 1.0 | 0.5-2.2 | | |
| | Meta-analysis model | 2009 | Outcome | 12 cases, 24 controls Studies included | Meta-RR | 95% CI | l ² | n |
| | Model 1 | | | | | | 0.0% | P _{heterogeneity} 0.84 |
| | | | Non-Hodgkin lymphoma | 1a, 2, 3, 4, 6, 7 | 1.3 | 1.0–1.6 | 0.0% | 0.84 |
| | Model 2 | | Non-Hodgkin lymphoma | 1b, 2, 3, 4, 6, 7 | 1.3 | 1.0–1.6 | | 0.59 |
| | Model 3 | | Non-Hodgkin lymphoma | 1a, 2, 3, 4, 5, 7 | 1.3 | 1.0–1.7 | 0.0% | |
| | Model 4 | | Non-Hodgkin lymphoma | 1b, 2, 3, 4, 5, 7 | 1.4 | 1.0–1.8 | 0.0% | 0.63 |
| | Eriksson et al.[14] | 2008 | B-cell lymphoma | Not reported | 1.87 | 0.998-3.51 | | |
| 3 | Cocco et al.[18] | 2013 | B-cell lymphoma | 4 cases, 2 controls | 3.1 | 0.6-17.1 | | |
| | Meta-analysis model | | Outcome | Studies included | Meta-RR | 95% CI | l ² | P _{heterogeneity} |
| | Model 1 | | B-cell lymphoma | 3, 8 | 2.0 | 1.1–3.6 | 0.0% | 0.58 |
| ; | Eriksson et al.[14] | 2008 | Diffuse large B-cell lymphoma | Not reported | 1.22 | 0.44-3.35 | | |
| 7 | Orsi et al. [17] | 2009 | Diffuse large B-cell lymphoma | 5 cases, 24 controls | 1.0 | 0.3–2.7 | | |
| ′ | Meta-analysis model | 2007 | Outcome | Studies included | Meta-RR | 95% CI | l ² | P _{heterogeneity} |
| | Model 1 | | Diffuse large B-cell lymphoma | 3, 7 | 1.1 | 0.5–2.3 | 0.0% | heterogeneity 0.79 |
| | Model 1 | | Diriuse large b-cell lymphoma | 5, 7 | 1.1 | 0.5-2.5 | 0.070 | 0.73 |
| 3 | Eriksson et al.[14] | 2008 | CLL/SLL | Not reported | 3.35 | 1.42-7.89 | | |
| 7 | Orsi et al. ^[17] | 2009 | CLL/SLL | 2 cases, 18 controls | 0.4 | 0.1-1.8 | | |
| | Meta-analysis model | | Outcome | Studies included | Meta-RR | 95% CI | l ² | Pheterogeneity |
| | Model 1, random effects | | CLL/SLL | 3, 7 | 1.3 | 0.2-10.0 | 83.7% | 0.01 |
| | Model 1, fixed effects | | CLL/SLL | 3, 7 | 1.9 | 0.9-4.0 | | |
| 3 | Eriksson et al.[14] | 2008 | Follicular lymphoma | Not reported | 1.89 | 0.62-5.79 | - | |
| 7 | Orsi et al.[17] | 2009 | Follicular lymphoma | 3 cases, 24 controls | 1.4 | 0.4–5.2 | | |
| | Meta-analysis model | 2007 | Outcome | Studies included | Meta-RR | 95% CI | l ² | P _{heterogeneity} |
| | Model 1 | | Follicular lymphoma | 3, 7 | 1.7 | 0.7–3.9 | 0.0% | 0.73 |
| , | Orsi et al.[17] | 2009 | Hairy-cell leukemia | 2 cases, 18 controls | 1.8 | 0.3–9.3 | | |
| | | | | | | | | |
| 9 | Nordström et al. [30] | 1998 | Hairy-cell leukemia | 4 cases, 5 controls | 3.1 M PD | 0.8–12 | l ² | |
| | Meta-analysis model | | Outcome | Studies included | Meta-RR | 95% CI | | P _{heterogeneity} |
| | Model 1 | | Hairy-cell leukemia | 7, 9 | 2.5 | 0.9–7.3 | 0.0% | 0.63 |
| , | Orsi et al.[17] | 2009 | Hodgkin lymphoma | 6 cases, 24 controls | 1.7 | 0.6-5.0 | | |
| 0 | Karunanayake et al.[31] | 2012 | Hodgkin lymphoma | 38 cases, 133 controls | 0.99 | 0.62-1.56 | | |
| | Meta-analysis model | | Outcome | Studies included | Meta-RR | 95% CI | l ² | Pheterogeneity |
| | Model 1 | | Hodgkin lymphoma | 7, 10 | 1.1 | 0.7–1.6 | 0.0% | 0.36 |
| ! | De Roos et al.[12] | 2005 | Multiple myeloma | 19 cases [†] | 2.6 | 0.7–9.4 | | |
| 7 | Orsi et al. [17] | 2003 | Multiple myeloma | 5 cases, 24 controls | 2.4 | 0.8-7.3 | | |
| | | 1993 | Multiple myeloma | 11 cases, 40 controls | 1.7 | 0.6-7.3 | | |
| | Brown et al [32] | | MINIMALE HIVEIUHIA | | | | | |
| 11 | Brown et al. [32] | | | 37 cases 121 controls | | | | |
| 12 | Kachuri et al.[33] | 2013 | Multiple myeloma | 32 cases, 121 controls | a. 1.19 (with proxies)b. 1.11 (without proxies) | a. 0.76–1.87 (with proxies) b. 0.66–1.86 (without proxies) | | |
| 12 | Kachuri et al. ^[33] Pahwa et al. ^[34] | 2013 2012 | Multiple myeloma Multiple myeloma | 32 cases, 121 controls 32 cases, 133 controls | | | | |
| 1 2 3 | Kachuri et al. ^[33] Pahwa et al. ^[34] Sorahan ^[26] | 2013 | Multiple myeloma | · | b. 1.11 (without proxies) | b. 0.66-1.86 (without proxies) | | |
| 11 12 13 | Kachuri et al. ^[33] Pahwa et al. ^[34] Sorahan ^[26] | 2013 2012 | Multiple myeloma Multiple myeloma | 32 cases, 133 controls | b. 1.11 (without proxies) 1.22 | b. 0.66–1.86 (without proxies) 0.77–1.93 | l ² | Pheterogeneity |
| 11 12 13 | Kachuri et al. ^[33] Pahwa et al. ^[34] | 2013 2012 | Multiple myeloma Multiple myeloma Multiple myeloma | 32 cases, 133 controls 24 cases | b. 1.11 (without proxies) 1.22 1.24 | b. 0.66–1.86 (without proxies) 0.77–1.93 0.52–2.94 | l² 0.0% | P _{heterogeneity} |
| 1 2 3 | Kachuri et al. ^[33] Pahwa et al. ^[34] Sorahan ^[26] Meta-analysis model | 2013 2012 | Multiple myeloma Multiple myeloma Multiple myeloma Outcome | 32 cases, 133 controls 24 cases Studies included | b. 1.11 (without proxies) 1.22 1.24 Meta-RR | b. 0.66–1.86 (without proxies) 0.77–1.93 0.52–2.94 95 % CI | - | P _{heterogeneity} 0.63 0.48 |
| 11 12 13 | Kachuri et al. ^[33] Pahwa et al. ^[34] Sorahan ^[26] Meta-analysis model Model 1 | 2013 2012 | Multiple myeloma Multiple myeloma Multiple myeloma Outcome Multiple myeloma | 32 cases, 133 controls 24 cases Studies included 7, 11, 12a, 14 | b. 1.11 (without proxies) 1.22 1.24 Meta-RR 1.4 | b. 0.66–1.86 (without proxies) 0.77–1.93 0.52–2.94 95% CI 1.0–1.9 | 0.0% | 0.63 |
| 1 2 3 | Kachuri et al. ^[33] Pahwa et al. ^[34] Sorahan ^[26] Meta-analysis model Model 1 Model 2 | 2013 2012 | Multiple myeloma Multiple myeloma Multiple myeloma Outcome Multiple myeloma Multiple myeloma | 32 cases, 133 controls 24 cases Studies included 7, 11, 12a, 14 2, 7, 11, 12a 7, 11, 12b, 14 | b. 1.11 (without proxies) 1.22 1.24 Meta-RR 1.4 1.5 | b. 0.66–1.86 (without proxies) 0.77–1.93 0.52–2.94 95% CI 1.0–1.9 1.0–2.1 | 0.0% | 0.63 0.48 |
| 11 12 13 | Rachuri et al. ^[33] Pahwa et al. ^[34] Sorahan ^[26] Meta-analysis model Model 1 Model 2 Model 3 | 2013 2012 | Multiple myeloma Multiple myeloma Multiple myeloma Outcome Multiple myeloma Multiple myeloma Multiple myeloma | 32 cases, 133 controls 24 cases Studies included 7, 11, 12a, 14 2, 7, 11, 12a | b. 1.11 (without proxies) 1.22 1.24 Meta-RR 1.4 1.5 1.4 | b. 0.66–1.86 (without proxies) 0.77–1.93 0.52–2.94 95% CI 1.0–1.9 1.0–2.1 0.9–1.9 | 0.0% 0.0% 0.0% | 0.63 0.48 0.58 |
| 11 12 13 14 | Rachuri et al. ^[33] Pahwa et al. ^[34] Sorahan ^[26] Meta-analysis model Model 1 Model 2 Model 3 Model 4 Model 5 | 2013 2012 2015 | Multiple myeloma Multiple myeloma Multiple myeloma Outcome Multiple myeloma | 32 cases, 133 controls 24 cases Studies included 7, 11, 12a, 14 2, 7, 11, 12a 7, 11, 12b, 14 7, 11, 13, 14 2, 7, 11, 13 | b. 1.11 (without proxies) 1.22 1.24 Meta-RR 1.4 1.5 1.4 1.4 1.5 | b. 0.66–1.86 (without proxies) 0.77–1.93 0.52–2.94 95% CI 1.0–1.9 1.0–2.1 0.9–1.9 1.0–2.0 1.0–2.1 | 0.0% 0.0% 0.0% 0.0% | 0.63 0.48 0.58 0.66 |
| 11 12 13 14 14 14 1 | Kachuri et al. ^[34] Pahwa et al. ^[34] Sorahan ^[26] Meta-analysis model Model 1 Model 2 Model 3 Model 4 Model 5 De Roos et al. ^[12] | 2013 2012 2015 2015 | Multiple myeloma Multiple myeloma Multiple myeloma Outcome Multiple myeloma Multiple myeloma Multiple myeloma Multiple myeloma Multiple myeloma Multiple myeloma Leukemia | 32 cases, 133 controls 24 cases Studies included 7, 11, 12a, 14 2, 7, 11, 12a 7, 11, 12b, 14 7, 11, 13, 14 2, 7, 11, 13 | b. 1.11 (without proxies) 1.22 1.24 Meta-RR 1.4 1.5 1.4 1.5 1.4 1.0 | b. 0.66–1.86 (without proxies) 0.77–1.93 0.52–2.94 95% CI 1.0–1.9 1.0–2.1 0.9–1.9 1.0–2.0 1.0–2.1 | 0.0% 0.0% 0.0% 0.0% | 0.63 0.48 0.58 0.66 |
| 11 12 13 14 14 | Kachuri et al. ^[34] Pahwa et al. ^[34] Sorahan ^[26] Meta-analysis model Model 1 Model 2 Model 3 Model 4 Model 5 De Roos et al. ^[12] Brown et al. ^[35] | 2013 2012 2015 2015 | Multiple myeloma Multiple myeloma Multiple myeloma Outcome Multiple myeloma Multiple myeloma Multiple myeloma Multiple myeloma Multiple myeloma Multiple myeloma Leukemia Leukemia | 32 cases, 133 controls 24 cases Studies included 7, 11, 12a, 14 2, 7, 11, 12b, 14 7, 11, 13, 14 2, 7, 11, 13 | b. 1.11 (without proxies) 1.22 1.24 Meta-RR 1.4 1.5 1.4 1.5 1.4 1.5 | b. 0.66–1.86 (without proxies) 0.77–1.93 0.52–2.94 95% CI 1.0–1.9 1.0–2.1 0.9–1.9 1.0–2.0 1.0–2.1 | 0.0% 0.0% 0.0% 0.0% | 0.63 0.48 0.58 0.66 |
| 11 12 13 14 | Kachuri et al. ^[34] Pahwa et al. ^[34] Sorahan ^[26] Meta-analysis model Model 1 Model 2 Model 3 Model 4 Model 5 De Roos et al. ^[12] | 2013 2012 2015 2015 | Multiple myeloma Multiple myeloma Multiple myeloma Outcome Multiple myeloma Multiple myeloma Multiple myeloma Multiple myeloma Multiple myeloma Multiple myeloma Leukemia | 32 cases, 133 controls 24 cases Studies included 7, 11, 12a, 14 2, 7, 11, 12a 7, 11, 12b, 14 7, 11, 13, 14 2, 7, 11, 13 | b. 1.11 (without proxies) 1.22 1.24 Meta-RR 1.4 1.5 1.4 1.5 1.4 1.0 | b. 0.66–1.86 (without proxies) 0.77–1.93 0.52–2.94 95% CI 1.0–1.9 1.0–2.1 0.9–1.9 1.0–2.0 1.0–2.1 | 0.0% 0.0% 0.0% 0.0% | 0.63 0.48 0.58 0.66 |

*Number of exposed cases is provided for the total cohort of 54,315 subjects; the number of exposed cases in the analytic cohort of 49,211 subjects is not stated.

†Number of exposed cases is provided for the analytic cohort of 40,719 subjects, as reported by Sorahan. [26]

Change does interval Cluster and the provided for the analytic cohort of 40,719 subjects, as reported by Sorahan. [26]

Cl: confidence interval; CLL: chronic lymphocytic leukemia; RR: relative risk; SLL: small lymphocytic lymphoma.

leukemia, one with unspecified B-cell lymphoma, and one with MM. Eriksson et al. [14] did not report the number of exposed cases, but overall the B-cell lymphomas in their study comprised 29% diffuse large B-cell lymphoma, 24% chronic lymphocytic leukemia/small lymphocytic lymphoma, 20% follicular lymphoma grades I–III, 16% other specified B-cell lymphoma, and 11% unspecified B-cell lymphoma; MM cases were not included.

The meta-RR for the association between any use of glyphosate and risk of diffuse large B-cell lymphoma, based on two studies, [14,17] was 1.1 (95% CI = 0.5–2.3) using both the random-effects and the fixed-effects models ($I^2 = 0.0\%$, $P_{heterogeneity} = 0.79$) (Table 3).

Based on the same two studies, [14,17] the meta-RR for the association between any use of glyphosate and risk of chronic lymphocytic leukemia/small lymphocytic lymphoma was 1.3 (95% CI = 0.2–10.0) according to the random-effects model and 1.9 (95% CI = 0.9–4.0) according to the fixed-effects model, with significant heterogeneity between the two included estimates ($I^2 = 83.7\%$, $p_{\text{heterogeneity}} = 0.01$) (Table 3).

Results for follicular lymphoma from these two studies, $^{[14,17]}$ by contrast, were not significantly heterogeneous ($I^2 = 0.0\%$, $p_{\text{heterogeneity}} = 0.73$), with a meta-RR of 1.7 (95% CI = 0.7–3.9) in both the random-effects and the fixed-effects models (Table 3).

| Authors | Year | RR | 95% CI | 1 | 1 - | - 1 | Relative weight (%) |
|---------------------|------|------|-------------|-----|----------------|-----|---------------------|
| Brown et al. [32] | 1993 | 1.7 | 0.8-3.6 | | | | 20.0 |
| Kachuri et al. [33] | 2013 | 1.19 | 0.76 - 1.87 | | - | | 55.7 |
| Orsi et al. [17] | 2009 | 2.4 | 0.8-7.3 | | - | - | 9.2 |
| Sorahan [26] | 2015 | 1.24 | 0.52 - 2.94 | | - | | 15.1 |
| Meta-RR | | 1.4 | 1.0-1.9 | | • | | |
| | | | | 0.1 | 1.0 | 10 | |

Figure 2. Forest plots of relative risk (RR) estimates and 95% confidence intervals (Cls) for the association between glyphosate exposure and risk of multiple myeloma. Meta-RRs were identical in random-effects and fixed-effects models.

Finally, the two studies that reported associations between any glyphosate use and risk of hairy-cell leukemia [17,30] yielded a meta-RR of 2.5 (95% CI = 0.9-7.3) in the random-effects and fixed-effects models ($I^2 = 0.0\%$, $p_{heterogeneity} = 0.63$) (Table 3).

HL

Both of the published, fully adjusted RRs and 95% CIs for the association between any glyphosate use and HL risk (Table 2) were included in the meta-analysis (Table 3). Based on two studies, [17,31] the meta-RR was 1.1 (95% CI = 0.7-1.6) in both the random-effects and the fixed-effects models, with $I^2 = 0.0\%$ and p_{heterogeneity} = 0.36 (Table 3). Publication bias was not evaluated due to the availability of only two studies of HL.

MM

All relevant RRs and 95% CIs for the association between glyphosate use and risk of MM, including estimates that did not contribute to the meta-analysis, are shown in Table 2. The independent estimates selected for inclusion in the meta-analysis are shown in Table 3.

The combined meta-RR for the association between any glyphosate use and risk of MM, based on four studies, [17,26,32,33] was 1.4 (95% CI = 1.0-1.9) according to the random-effects and fixed-effects models (Table 3, Fig. 2). On the basis of the I^2 value of 0.0% and the P-value of 0.63 for Cochran's O statistic, between-study heterogeneity was not evident. Egger's linear regression approach yielded no significant evidence of publication bias (one-tailed P-value for asymmetry = 0.10), while the imputed meta-RR using the trim-and-fill procedure to adjust for publication bias was 1.3 (95% CI = 0.9-1.8).

Several secondary analyses were conducted for MM by replacing RRs in the primary meta-analysis with alternative estimates (Table 3). When the RR reported by De Roos et al., [12] who excluded cohort members with missing data from their analysis, was substituted for the one reported by Sorahan, [26] who included such subjects by creating a separate category for missing or unknown data, the meta-RR was slightly increased to 1.5 (95% CI = 1.0-2.1) and was the same for random-effects and fixed-effects models. When the main RR from Kachuri et al.[33] was replaced with the RR from the same study after exclusion of data reported by proxy respondents, the meta-RR was not appreciably different from the original estimate (alternative meta-RR = 1.4, 95% CI = 0.9-1.9 in random-effects and fixed-effects models). Another secondary analysis included the RR reported by Pahwa et al., $^{[34]}$ who adjusted for a slightly different (and smaller) set of confounders than Kachuri et al.[33] and also retained controls who were too young to have any agematched MM cases in this Canadian study. This change had

minimal impact on the meta-RR (1.4, 95% CI = 1.0-2.0; same for random-effects and fixed-effects models). When both the De Roos et al. [12] and the Pahwa et al. [34] substitutions were made, the resultant meta-RR was the same as that when only De Roos et al. [12] was used (meta-RR = 1.5, 95% CI = 1.0-1.2 in random-effects and fixed-effects models).

Leukemia

Of the four published RRs and 95% CIs for the association between any use of glyphosate and risk of leukemia (Table 2), three (excluding one age-adjusted RR in favor of a more fully adjusted RR from De Roos et al. [12]) were included in the metaanalysis (Table 3). The meta-RR based on three studies^[12,35,36] was 1.0 (95% CI = 0.6-1.5) using the random-effects model and the fixed-effects model ($I^2 = 0.0\%$, $p_{heterogeneity} = 0.92$) (Table 3). Publication bias was not assessed because only three studies of leukemia were available.

Sensitivity analysis

A sensitivity analysis was conducted for overall NHL only (Table 4), because other outcomes had an insufficient number of studies for stratification. In all strata, the randomeffects and fixed-effects meta-RRs were identical and I^2 was 0.0%. Results did not differ substantially from the main meta-RR (1.3, 95% CI = 1.0-1.6) when the analysis was restricted to case-control studies (meta-RR = 1.3, 95% CI = 1.0-1.7) or those with population-based controls (meta-RR = 1.4, 95% CI = 1.0-1.8). Meta-analysis could not be conducted for cohort studies or studies with hospital-based

Table 4. Sensitivity analysis of the association between glyphosate exposure and risk of non-Hodgkin lymphoma (NHL).

| Stratum | Number of studies | Meta-RR* | 95% CI |
|---------------------|-------------------|----------|---------|
| All | 6 | 1.3 | 1.0-1.6 |
| Case-control | 5 | 1.3 | 1.0–1.7 |
| Cohort | 1 | NR | |
| Population controls | 4 | 1.4 | 1.0-1.8 |
| Hospital controls | 1 | NR | |
| Males only | 4 2 | 1.3 | 1.0-1.7 |
| Males and females | | 1.2 | 0.8-1.8 |
| North America | 3 | 1.2 | 1.0-1.6 |
| Europe | 3 | 1.3 | 0.8-2.1 |
| Sweden | 2 | 1.6 | 0.9-2.8 |
| Cases in 1980s | 2 | 1.6 | 1.0-2.7 |
| Cases in 1990s | 4 | 1.2 | 1.0-1.6 |
| Cases in 2000s | 3 | 1.2 | 0.8-1.7 |

^{*}All meta-RRs were identical in random-effects and fixed-effects models. CI: confidence interval; meta-RR: meta-analysis relative risk; NR: not reported, when only one study was available.



controls because only one of each of these study types was available. No major differences were detected between studies restricted to males (meta-RR = 1.3, 95% CI = 1.0-1.7) and those that included males and females (meta-RR = 1.2, 95% CI = 0.8-1.8) or between those conducted in North America (meta-RR = 1.2, 95% CI = 1.0-1.6) and those conducted in Europe (meta-RR = 1.3, 95% CI = 0.8-2.1). Prompted by Schinasi and Leon, [11] we also conducted a stratified meta-analysis of the two studies conducted in Sweden[14,15] and found a stronger, albeit statistically non-significant, association in these particular studies (meta-RR = 1.6, 95% CI = 0.9-2.8). The estimated meta-RR declined somewhat from studies that ascertained cases in the 1980s (meta-RR = 1.6, 95% CI = 1.0-2.7) to those conducted in the 1990s (meta-RR = 1.2, 95% CI = 1.0-1.6) to those conducted in the 2000s (meta-RR = 1.2, 95% CI = 0.8-1.7).

Exposure-response trends

NHL and subtypes. Three studies evaluated exposure-response trends between glyphosate use and NHL risk, with exposure classified as cumulative lifetime^[12,14] or annual^[16] days of glyphosate use (Table 2). Two studies detected some evidence of a positive exposure-response trend (statistical significance not reported), [14,16] whereas the other did not. [12] All of these studies relied wholly or in part on evaluating days of glyphosate use in an attempt to quantify exposure; however, this metric has been shown to be a poor indicator of actual glyphosate dose received.^[52]

In a model adjusted for age, sex, and year of diagnosis or enrollment, Eriksson et al. [14] found that the RR of NHL was higher with > 10 days of lifetime glyphosate use (RR = 2.36, 95% CI = 1.04-5.37) than with \leq 10 days (RR = 1.69, 95% CI = 0.70-4.07), compared with no pesticide use. Also, the RR of NHL was higher after more than 10 years since first use of glyphosate (RR = 2.26, 95% CI = 1.16-4.40) than after 1-10 years (RR = 1.11, 95% CI = 0.24-5.08). Statistical tests for trend were not performed, and exposure-response analyses adjusted for other potential confounders (i.e., 2-methyl-4chlorophenoxyacetic acid (MCPA), 2,4,5-trichlorophenoxyacetic acid (2,4,5-T) and/or 2,4-dichlorophenoxyacetic acid (2,4-D), mercurial seed dressing, arsenic, creosote, and tar) were not presented, even though adjustment for these characteristics attenuated the RR for overall glyphosate use from 2.02 to 1.51.

McDuffie et al.[16] reported that the RR for more than two days of glyphosate use per year (RR = 2.12, 95% CI = 1.20-3.73) was higher than that for up to two days per year (RR =1.00, 95% CI = 0.63-1.57), compared with never use, adjusting for age and province of residence. Tests for a significant exposure-response trend were not performed, and results were not reported after adjustment for other potential confounders (i.e., personal medical history and family history of cancer; adjustment for these characteristics attenuated the RR for overall glyphosate use from 1.26 to 1.20) or significantly associated pesticides (i.e., aldrin, dicamba, and mecoprop) in this study population.

The most detailed analysis of glyphosate-NHL exposureresponse trends was performed by De Roos et al., [12] who examined tertiles of cumulative lifetime days of glyphosate use (1-20, 21-56, or 57-2,678 days) and tertiles of intensity-weighted cumulative days of use (i.e., years of use x days per year x intensity level, where intensity was defined as (mixing status + application method + equipment repair status) × personal protective equipment use). In analyses adjusted for age, education, smoking, alcohol, family history of cancer, and state of residence, no significant trend was detected for NHL risk in association with increasing cumulative days of glyphosate use (RRs for tertiles 1, 2, and 3, respectively = 1.0 (referent), 0.7(95% CI = 0.4–1.4), and 0.9 (95% CI = 0.5–1.6); $p_{trend} = 0.73$) or intensity-weighted cumulative exposure days (RRs = 1.0 (referent), 0.6 (95% CI = 0.3-1.1), and 0.8 (95% CI = 0.5-1.4); $p_{trend} = 0.99$).

Exposure-response trends between glyphosate use and risk of specific NHL subtypes were not evaluated in any of the included studies.

HL. No studies assessed exposure-response trends between glyphosate use and risk of HL.

MM. Three studies reported exposure-response trends between glyphosate use and MM risk, including the two analyses based on the same Agricultural Health Study cohort dataset^[12,26] and the Canadian case-control study^[33] (Table 2). The case-control study found mixed evidence of a positive trend (statistical significance not reported), while a positive trend was detected in one analysis of the cohort data^[12] but not the other.[25]

The Canadian case-control study found a lower risk of MM among those who used glyphosate for up to two days per year than those who had never used glyphosate (RR = 0.72, 95% CI = 0.39-1.32). However, risk was higher in those with more than two days of glyphosate use per year (RR = 2.04, 95%CI = 0.98-4.23), adjusting for age, province of residence, proxy status, smoking, personal medical history, and family history of cancer. Results were similar after exclusion of data reported by proxy subjects. The authors did not conduct statistical tests for exposure-response trends.

Based on the 55% of Agricultural Health Study cohort members who had available exposure and covariate data, De Roos et al. [12] reported a positive, albeit statistically non-significant, trend between MM risk and increasing tertiles of cumulative days of glyphosate use (RRs for tertiles 1, 2, and 3, respectively = 1.0 (referent), 1.1 (95% CI = 0.4-3.5), and 1.9 (95% CI = 0.6-6.3); $p_{trend} = 0.27$) or intensity-weighted cumulative days of use (RRs = 1.0 (referent), 1.2 (95% CI = 0.4-3.8), and 2.1(95% CI = 0.6-7.0); $p_{trend} = 0.17$). These estimates were adjusted for age, education, smoking, alcohol, family history of cancer, state of residence, the five pesticides for which cumulative-use variables were most highly associated with glyphosate cumulative use days (i.e., 2,4-D, alachlor, atrazine, metolachlor, and trifluralin), and the five pesticides that were most highly associated with ever use of glyphosate (i.e., benomyl, maneb, paraquat, carbaryl, and diazinon). When intensity alone was analyzed in association with MM risk, the RR for the highest versus the lowest tertile was 0.6 (95% CI = 0.2-1.8), indicating that the suggested trend was due only to total days of use. When subjects who never used glyphosate were set as the reference group, the RRs for tertiles 1, 2, and 3 of cumulative days



of use were 2.3 (95% CI = 0.6-8.9), 2.6 (95% CI = 0.6-11.5), and 4.4 (95% CI = 1.0-20.2); $p_{trend} = 0.09$. When cumulative use was categorized into quartiles, the RR for the highest quartile versus never use was 6.6 (95% CI = 1.4-30.6); $p_{trend} = 0.01$.

In contrast to De Roos et al., [12] Sorahan [26] included more than 53,000 eligible cohort members in the analysis (excluding only those with a history of cancer before enrollment, loss to follow-up, missing data on age at enrollment, or missing data on glyphosate use) by creating separate categories for missing or unknown exposure and covariate data. Adjusting for age, sex, education, smoking, alcohol, family history of cancer, and the same 10 pesticides as De Roos et al., [12] the RRs for each tertile of cumulative days of glyphosate use, compared with never use, were 1.14 (95% CI = 0.43-3.03), 1.52 (95% CI =0.54-4.34), and 1.38 (95% CI = 0.42-4.45); $p_{trend} = 0.48$ using category scores of 1-4, p_{trend} > 0.50 using mean exposures within categories. RRs for increasing tertiles of intensityweighted days of use versus never use were 1.00 (95% CI = 0.33-3.00), 1.27 (95% CI = 0.45-3.56), and 1.87 (95% CI = 0.67–5.27); $p_{trend}=0.22$ using scores, $p_{trend}=0.18$ using means. When Sorahan^[26] expanded the eligible cohort to 55,934 subjects to include those with unknown use of glyphosate, he again detected no significant exposure-response trends with respect to either cumulative days of use (for tertiles 1, 2, and 3 and unknown use versus never use, respectively, RRs = 1.11 (95% CI = 0.44-2.83), 1.45 (95% CI = 0.54-3.88), 1.17(95% CI = 0.40-3.41), and 1.19 (95% CI = 0.25-5.65); p_{trend} > 0.50 across categories of known use using scores or means, excluding unknown) or intensity-weighted cumulative days of use (RRs = 0.95 (95% CI = 0.33-2.75), 1.19 (95% CI = 0.44-3.19), 1.58 (95% CI = 0.62-4.05), and 1.04 (95% CI = 0.22-4.92); $p_{trend} = 0.30$ using scores, $p_{trend} = 0.26$ using means, excluding unknown).

Leukemia. The De Roos et al. [12] study based on the Agricultural Health Study cohort was the only study that reported exposure-response trends between glyphosate use and risk of leukemia (Table 2). No significant trend was observed between increasing tertiles of cumulative days of glyphosate use (RRs = 1.0 (referent), 1.9 (95% CI = 0.8-4.5), and 1.0 (95% CI = 0.4-2.9) for tertiles 1, 2, and 3, respectively; $p_{trend} = 0.61$) or intensity-weighted cumulative days of use (RRs = 1.0 (referent), 1.9 (95% CI = 0.8-4.7), and 0.7 (95% CI = 0.2-2.1); $p_{trend} = 0.11$), adjusting for demographic and lifestyle factors as well as other pesticides.

Evaluation of bias

Selection bias

All studies of the association between glyphosate exposure and risk of LHC were case-control studies except for the Agricultural Health Study, the prospective cohort study that served as the basis for the studies by De Roos et al. [12] and Sorahan. [26] In case-control studies, differences in participation patterns between cases and controls can result in selection bias if participation is related to the exposure of interest. In cohort studies, selection bias can occur if loss to follow-up is related to the exposure and outcome of interest or, less commonly, if baseline participation differs by exposure status and risk of developing the outcome of interest in the future (e.g., based on having a positive family history of an outcome with a genetic susceptibility component). Selection bias in any study also can occur if inclusion in the data analysis, e.g., predicated on data completeness, differs by exposure and outcome status. In general, lower participation, follow-up, or data completeness and large differences in participation between groups increase the potential magnitude of selection bias.

Table 1 shows the reported participation and follow-up proportions in all reviewed studies. Most studies did not report data completeness. The substantial differences in participation between cases and controls in the European multi-center study, [18] the most recent Swedish study, [14] and the Canadian study, which also had relatively low absolute participation proportions of <70% for cases and <50% for controls, [16,28,31,33,34] are of particular concern. However, the smaller discrepancies between case and control participation in other studies also could have produced selection bias. Moreover, even identical participation by cases and controls can obscure differences in reasons for study participation that could result in bias.

Given that several case-control studies were originally designed to evaluate associations between pesticides and risk of LHC, [13-16,28,31-35] it is plausible that cases with a history of agricultural pesticide use were more likely than controls to participate, thereby biasing results toward a positive association for glyphosate as well as other pesticides. It is also possible that certain sources of controls in some of these studies (e.g., residential telephone calls and voter lists) were more likely to identify individuals who were not farmers, again biasing results toward a positive association. Investigators from the Canadian study^[16,28,31,33,34] reported that an analysis of postal codes showed that respondents and non-respondents did not differ significantly in terms of rural versus urban residence, but they could not examine differences in occupation or pesticide use.

Although the initial follow-up completion of >99% in the Agricultural Health Study was high, [12,25] the sizeable proportions of subjects with missing data raise concerns about selection bias. Specifically, 88% of the eligible cohort (excluding those who were diagnosed with cancer before enrollment or were lost to follow-up) provided usable data on ever use of glyphosate and key demographic and lifestyle covariates, 73% additionally provided data on use of other pesticides, 65-66% contributed to analyses of cumulative days of glyphosate use (with or without intensity weighting), and 55% contributed to analyses of cumulative use additionally adjusted for other pesticides. Questionnaire completion could conceivably have varied by demographic and lifestyle factors that are associated with LHC risk, thereby producing bias. Neither analysis accounted for missing data using methods such as multiple imputation or inverse probability weighting.

Differential data completeness by disease status is more likely to occur in case-control studies, such as the pooled Midwestern U.S. study conducted by De Roos et al. [13] In this study, the analysis of multiple pesticides excluded 25% of cases and 25% of controls who lacked complete data. Although the overall frequency of missing data was the same between cases and controls, this exclusion could have led to selection bias if subjects'



reasons for providing complete data, and thus being included in the analysis, differed by disease status and were related to glyphosate exposure status. The authors also excluded subjects who had lived or worked on a farm before age 18 years. If glyphosate use was more common in such subjects, then RR estimates would have been biased upward if a childhood farm environment was inversely associated with NHL risk^[53] and biased downward if the association was positive. [54]

Exposure misclassification

All of the included studies assessed use of glyphosate and other pesticides based on self-reported information (Table 1), which is prone to various types of error, such as better recall by cases than controls and by subjects than proxies, inaccurate recall of specific pesticides and amounts used, and a lack of the best measure of biological dose received. Thus, probable exposure misclassification is a key limitation of all of these studies. The degree of misclassification may vary by mode of data collection, for example, by written questionnaire, telephone interview, or in-person interview. [56] The extent of misclassification also may depend on questionnaire structure, for example, whether subjects were asked in an open-ended manner to report use of any pesticides or whether they were prompted to report use of specific pesticides based on a prepared list. [57] Some authors did not clearly describe the structure of their study's questions on pesticide use.

Of the eight independent study populations included in this review (seven studies of NHL with or without other types of LHC and one study of leukemia), three provided information on validation of their exposure assessment methods: the Canadian case-control study, [16,28,31,33,34] the Agricultural Health Study, [12,26] and the Kansas case-control study [47] that contributed to the pooled Midwestern U.S. study by De Roos et al. [13] Overall, these studies do not establish the validity of selfreported information on glyphosate use; rather, the limited results suggest considerable error and inconsistency in such data.

Specifically, in the Canadian study, Dosman et al. [58] reported on the results of a validation pilot study of 21 volunteer farmers whose self-reported pesticide use was compared with written records of pesticide purchases through their local agrochemical supplier. Of the 21 farmers, 17 (81%) had a supplier who had retained written records; the remaining four transactions were conducted with cash. Based on the written records, 146 (65%) of 226 chemicals reported by farmers were verified; 50 of the unverified reports were potentially explained by aerial applications, home and garden use, use more than five years in the past (i.e., during 1958–1984), or use outside of Canada. In 32 instances (for 25 chemicals) the suppliers' records indicated a purchase of chemicals that was unreported by the farmer; 2 of these were for glyphosate. Detailed self-reported exposure (e.g., frequency, intensity, and duration of use of specific pesticides) could not be validated in this pilot study.

Likewise, Hoar et al. [47] reported that suppliers for 110 subjects in the Kansas study (out of 130 sought) were located and provided information on the subjects' crops and herbicide and insecticide purchases as "corroborative evidence" of selfreported pesticide use. The authors observed that suppliers usually reported less pesticide use than subjects; that agreement on specific years of use was better for insecticide use than herbicide use; that the differences between agreement for cases and controls were not consistent; and that agreement between suppliers and subjects was better for pesticide use within the last 10 years than for earlier use. Quantitative results on concordance were not provided by Hoar et al., [47] but in a summary of this study shared with Dosman et al. [58] the authors stated that reports on herbicide use agreed 59% of the time, with little variation by crop type, and that reports on insecticide use also agreed 59% of the time, but differed by crop type.

In the Agricultural Health Study, the reliability of the question on ever having mixed or applied glyphosate was evaluated by comparing responses to two questionnaires completed one year apart by 3,763 pesticide applicators. [59] Agreement on a positive response to the question was 82%, and the kappa statistic value for inter-rater agreement was moderate (0.54, 95% CI = 0.52-0.58). For more detailed questions about glyphosate use, including years mixed or applied, days per year mixed or applied, and decade first applied, the percentage with exact agreement ranged from 52% to 62% and kappa ranged from 0.37 to 0.71. These metrics evaluated only the reliability (i.e., reproducibility) of self-reported glyphosate use, not its accuracy.

Subsequent exposure validation studies for other pesticides in the Agricultural Health Study, based on comparisons between exposure intensity estimated from an expert-derived algorithm using self-reported or directly observed exposure data and pesticide biomarker levels measured in urine, yielded Spearman correlation coefficients between 0.4 and 0.8, depending on the type of pesticide. [60,61] Correlations with urinary biomarker levels were poorer for self-reported determinants of pesticide exposure such as kilograms of active ingredient, hours spent mixing and applying, and number of acres treated, with correlation coefficients of -0.4 to 0.2, but application method and use of personal protective equipment were found to be important determinants of exposure intensity. However, the latter factors were evaluated in the study questionnaire only for pesticides or pesticide classes in general, not for glyphosate or other individual pesticides; [62] thus, limitations remain in the assessment of specific pesticide exposures.

Several studies included a sizeable proportion of surveys that were completed by proxy respondents for deceased or otherwise unavailable cases and controls (Table 1). The use of exposure data reported by surrogates most likely resulted in even poorer accuracy of exposure information in these studies. Although some exposure misclassification may have been nondifferential by disease status, such error does not inevitably result in underestimated exposure-disease associations unless additional strict conditions are met, such as independence from other classification errors. [63,64]

Furthermore, differential exposure misclassification in casecontrol studies can readily result in overestimated associations. Reasonable scenarios include more accurate and/or detailed recollection of past exposures by cases, who are more motivated than controls to try to understand the potential causes of their disease; false recollection by cases, who are more aware of scientific hypotheses or media reports that a certain exposure has been linked to their disease; and unconscious influence by study investigators who are aware of causal hypotheses and subjects' case-control status. Only the authors of the Swedish

studies, [14,15] the French study, [17] and the Nebraska component of the pooled Midwestern U.S. study^[48] specifically stated that investigators were blinded to case-control status. In reality, such blinding is often difficult to achieve in studies that collect interview data.

Others have discussed in detail the problems of estimating individual subjects' exposure to glyphosate from responses to interviews and questionnaires asking about days of use, mixing and application procedures, use of personal protective equipment, and other work practices. [19,52] Acquavella et al. [52] reported that any given day of pesticide use can entail highly variable amounts of pesticides used and numbers of mixing operations, and that urine concentrations of glyphosate were poorly correlated with lifetime average exposure intensity scores derived from data self-reported by farmers using this agent. Although recall bias between cases and controls generally might be anticipated to affect all specific pesticides (including glyphosate) equally, variation in the degree of misclassification due to these and other factors affecting usage and exposure could result in different pesticide-specific associations.

Most of the case-control studies did not use procedures to exclude glyphosate exposure that might have occurred after disease onset. The Swedish studies omitted glyphosate use within one year prior to diagnosis or the index date in controls, [15,30] or within the same calendar year or the year before. [14] In some cases, however, these restrictions may not have been sufficient to exclude exposure that occurred during the latency period between disease onset and diagnosis. Inclusion of any such post-disease exposure would have led to misclassification.

Finally, exposure misclassification resulting from the crude dichotomization of glyphosate use as ever versus never is an important limitation of most of the included studies. This classification conflates individuals with considerably different frequencies, intensities, and durations of glyphosate use, and precludes potentially informative analyses of any gradient in LHC risk with increasing glyphosate exposure. As described earlier in the section on exposure-response trends, only three independent studies reported on glyphosate use in more than two (ever vs. never) categories, and only the Agricultural Health Study evaluated more than three exposure categories.

Confounding

As shown in Table 1, the degree of control for confounding varied widely among the reviewed studies. Although several studies considered potential confounding by other pesticides or pesticide families, only a minority[12-15,26,28] reported RR estimates for the association between glyphosate use and LHC risk adjusted for use of other pesticides. Given that Schinasi and Leon^[11] found significant associations between NHL risk and several other types of pesticides, including carbamate insecticides, organophosphorus insecticides, lindane, and MCPA, and numerous other associations of specific pesticides with LHC risk have been reported in the literature (e.g., [65,66])—and because most people who use pesticides occupationally are exposed to multiple pesticides—it is important to control for confounding, whether direct or indirect (if pesticides are surrogates for other risk factors), by these agents.

None of the studies controlled for potential confounding by agricultural exposures other than pesticides, such as other

agricultural chemicals, farm animals, allergens, and infectious agents. These exposures have been hypothesized, and in some studies shown, to be associated with risk of NHL, HL, MM, or leukemia, [67-73] and they are probably correlated with glyphosate use, making them potential confounders of associations between glyphosate and LHC risk. Medical history, certain infections, diet, alcohol consumption, and obesity also may be associated with risk of these malignancies^[74-77] and could vary by glyphosate use, again making them possible confounders. Even in studies where numerous confounders were included in multivariable regression models, crude categorization or other misclassification of confounders could have enabled residual confounding of observed associations. The direction and magnitude of confounding depend on the relationships of each factor with glyphosate use and LHC risk, and are therefore difficult to predict.

Other issues

Additional issues related to the design, conduct, and reporting of the included studies also could have affected study results and their interpretation. For instance, Hardell et al. [15] enrolled some prevalent rather than incident cases, since eligible NHL cases were diagnosed in 1987-1990 but interviewed in 1993-1995. [27] The relatively long time interval between diagnosis and interview may have hampered recollection of past exposures, thereby undermining the accuracy of self-reported exposure data in this study. The delay between diagnosis and interview also almost certainly increased the proportion of cases and matched controls who were deceased (43%) and had proxy interviews, leading to further exposure misclassification.

In the studies by De Roos et al. [13] and Brown et al., [32,35] LHC cases were diagnosed in 1979-1986, 1980-1983, and 1980-1984, respectively. With glyphosate having come to market in 1974, the cases in these studies would have had a relatively short potential induction time since first use of glyphosate. However, few studies to date have considered the issue of induction time. The Agricultural Health Study collected information on decade of first use of glyphosate in the baseline questionnaire for private pesticide applicators, [62] but did not use this information in the published analysis. [12] If glyphosate is a cause of LHC, the actual induction time is unknown because the mechanism of carcinogenesis is not established.

Orsi et al.,[17] Kaufman et al.,[36] and four of the six study centers included in Cocco et al.^[18] enrolled hospital-based rather than population-based cases and controls. Given that farmers have lower hospitalization rates than non-farmers, [78] hospital-based controls may be less likely than populationbased controls to report agricultural occupational exposures, including pesticides, thereby resulting in overestimated RRs for pesticide use. On the other hand, occupational injuries are more common in agriculture than in general private industry, [79] possibly leading to oversampling of farmers from hospital trauma/emergency and orthopedics departments, which might result in underestimated RRs. We did not observe any meaningful change in the meta-RR after restriction to population-based case-control studies.

As noted in Table 1, many possible analyses were not conducted or not reported by authors. De Roos et al. [13] specifically acknowledged that they did not report results for pesticide combinations that were analyzed but yielded statistically null



associations for joint effects, and Hohenadel et al. [28] likewise did not show results for pesticide combinations without evidence of joint effects. Most other authors did not explicitly state when null results were not reported, but the Methods sections of several papers suggested that certain analyses were performed, yet not shown. Given the widespread predilection for emphasizing statistically significant associations in published research articles, [80] unreported results probably are usually statistically null. The omission of null results is a form of reporting bias that favors positive associations.

Other evidence suggests that statistically null associations between glyphosate and LHC risk have been underreported in the epidemiologic literature. For example, two of the studies that contributed to the pooled analysis conducted by De Roos et al. [13] apparently collected information on glyphosate use, yet associations between glyphosate and NHL risk were not reported in the original publications. [47,48] In an analysis of interactions between pesticide use and asthma, allergies, or hay fever diagnosis in relation to NHL risk in the Canadian casecontrol study,[81] results were reported for several specific pesticides, but not glyphosate, even though information was available for glyphosate use. The most probable scenario in each of these cases is that no significant association was detected between glyphosate use and NHL risk. The omission of such results from the published literature represents a distortion of the body of epidemiologic evidence.

The largest number of studies included in any of the metaanalyses described here was six (in the analysis of NHL), and the majority of meta-analyses (of HL, B-cell lymphoma, diffuse large B-cell lymphoma, chronic lymphocytic leukemia/small lymphocytic lymphoma, follicular lymphoma, and hairy-cell leukemia) included only two studies. The small number of available studies limits the robustness of the estimated meta-RRs, as well as the ability to perform informative sensitivity analysis and evaluation of heterogeneity and publication bias. Even with 10 contributing studies (which we lacked), the statistical power to detect modest heterogeneity using Cochran's Q statistic is "low." [42] The small number of studies also provides little opportunity to qualitatively investigate possible sources of heterogeneity by subject characteristics or study design. Thus, the results of the meta-analyses and related statistical tests reported here should be interpreted cautiously in light of the sparse and possibly selectively published literature, as well as the high potential for bias and confounding in most of the available studies.

Overall evaluation

The validity of the meta-RRs for glyphosate use and LHC risk reported here and by others^[11] is uncertain because systematic error due to bias and confounding cannot reasonably be ruled out as explanations for the observed associations (including both positive and null associations). In addition, an evaluation of the association between glyphosate exposure and risk of LHC based on the Bradford Hill viewpoints does not favor a causal relationship with NHL, any NHL subtype, HL, MM, or leukemia. These nine viewpoints are strength, consistency, specificity, temporality, biological gradient, plausibility, coherence, experiment, and analogy.

To evaluate the strength of the association between glyphosate use and risk of each type of LHC, we considered the magnitude of study-specific RRs and the corresponding meta-RRs. In individual studies, estimates of the association between glyphosate use and risk of NHL ranged between 1.0 and 2.1, and estimates of the association with NHL subtypes ranged between 0.4 and 3.35 (Table 3). For HL, the two estimates of association were 0.99 and 1.7. For MM, RRs ranged between 1.0 and 2.4, and those for leukemia ranged between 0.9 and 1.40. Most study-specific estimates were between 1.0 and 1.5. The estimated meta-RRs for all LHC outcomes, including those calculated in secondary and sensitivity analyses, ranged between 1.0 (for leukemia) and 2.5 (for hairy-cell leukemia). The meta-RRs calculated based on at least four studies ranged between 1.3 and 1.4. These associations are not of sufficient magnitude to exclude modest bias or confounding as reasonable explanations for the observed results.

Results were not consistent between case-control studies of NHL and the one prospective cohort study of NHL, which reported no association. [12] Even among the six studies that contributed to the meta-analysis of NHL, RR point estimates varied by more than two-fold, only one statistically significant positive association was observed, and results from some studies were internally inconsistent (Table 3). Another, arguably more appropriately adjusted RR (from a hierarchical regression model) that was 24% lower and statistically non-significant was reported in the same study that found a significant association. [13] The lack of statistically significant heterogeneity among studies of NHL, based on an underpowered statistical test, does not indicate consistency of results. For NHL subtypes, RR estimates also were variable, except for diffuse large B-cell lymphoma, for which both estimates were close to 1.0. Only one statistically significant positive association was detected (for chronic lymphocytic leukemia/small lymphocytic lymphoma), [14] and this result was contradicted by a non-significant inverse association in the other study of this outcome.^[17] No significant associations with ever use of glyphosate were detected for HL, MM, or leukemia, and for MM the RR point estimates varied by more than two-fold. Results for MM in the Agricultural Health Study were internally inconsistent; [12,26] and the positive association with cumulative glyphosate exposure probably was due largely to selection bias.

Numerous associations have been hypothesized between glyphosate exposure and diverse health outcomes, and between various exposures and risk of NHL, NHL subtypes, HL, MM, or leukemia. Thus, the putative associations are not specific to either the exposure or any of the outcomes. As noted by Bradford Hill, [46] "diseases may have more than one cause" and "one-to-one relationships are not frequent"; therefore, a lack of specificity does not detract from a causal hypothesis.

In case-control studies, where exposure assessment was retrospective, a temporal sequence was not definitively established with glyphosate use preceding the time of disease onset. Although some studies attempted to exclude use close to the time of case diagnosis (or enrollment, for controls),[14,15,30] in practice individuals may not accurately recall the timing of use. Only the prospective Agricultural Health Study[12,26] was designed to collect information on glyphosate use prior to cancer ascertainment. However, the authors did not exclude malignancies diagnosed close to



(e.g., within one year of) study enrollment, nor did they report the distribution of diagnoses with respect to time since first use of glyphosate. Thus, some preclinical cancers may have existed prior to study entry and, possibly, prior to at least some reported glyphosate use.

As discussed in detail earlier, in the three studies of NHL with information on frequency, intensity, and/or duration of glyphosate use, [12,14,16] a positive biological gradient was not consistently demonstrated and was notably lacking in the Agricultural Health Study, [12] which had the most detailed exposure information (Table 2). One case-control study^[33] and one prospective cohort study^[12] of MM reported results suggesting a positive biological gradient with glyphosate use, but the alternative analysis of the Agricultural Health Study data^[26] did not demonstrate such a trend. No data were available to evaluate exposure-response trends between glyphosate and risk of NHL subtypes or HL, and the single study with such data for leukemia found no apparent trend. [12]

Inhalation exposure to glyphosate from agricultural or residential uses is likely to be slight due to glyphosate's extremely low vapor pressure. [82] Although dermal contact can be considerable, the very low skin penetrability of glyphosate^[83] should result in minimal, if any, biologically absorbed dose. A study of farm families with a lower limit of detection of 0.001 μ g/mL (1 ppb) found that 40% of glyphosate applicators had undetectable urinary glyphosate, which reflects all routes of exposure (dermal, inhalation, and oral).^[84] Among those with detectable urinary glyphosate, the distribution of concentrations was right skewed, with a peak geometric mean concentration of $0.0032 \mu g/mL$ (3.2 ppb) on the day of application and declining thereafter. A review of seven human biomonitoring studies of glyphosate (including^[84]) yielded the conclusion that "no health concern was revealed because the resulting exposure estimates were by magnitudes lower" than the science-based acceptable daily intake and the acceptable operator exposure level proposed by EFSA. [85] Glyphosate is usually applied in agricultural operations only a few days per year. Given the low biological dose of glyphosate that is expected to be sustained, along with the lack of information on the mechanism of carcinogenesis that may exist in humans, the biological plausibility of LHC development due to typical glyphosate exposure has not been established.

IARC recently determined based on their process that there is "sufficient" evidence of carcinogenicity of glyphosate in experimental animals and mechanistic evidence of genotoxicity and oxidative stress. [6] By contrast, U.S. EPA, [86] JMPR, [3] BfR, [1] EFSA, [9] and others [87,88] concluded that glyphosate does not have genotoxic, mutagenic, or carcinogenic effects in in vivo animal and in vitro studies, and that the negative findings constitute evidence against carcinogenicity. Given these widely divergent opinions, one cannot unambiguously conclude whether the scientific evidence is coherent with the hypothesis that glyphosate causes any or all LHC.

No true experimental evidence exists regarding the association between glyphosate exposure and risk of LHC in humans. However, positive associations between farming and risk of LHC were detected prior to 1974, when glyphosate was first commercially marketed. [89,90] Thus, if the apparent associations between farming and risk of LHC are due to causal agricultural

exposures, they cannot be explained only by glyphosate exposure. Likewise, the recent worldwide increase (followed by a plateau or decline) in NHL incidence began before the 1970s^[91,92]—although any impact of glyphosate on NHL incidence trends might be obscured by stronger risk factors. No marked increase in the incidence of HL, MM, or leukemia has been observed in parallel with the introduction and expansion of glyphosate use. [93–96]

Finally, numerous analogies exist to support or oppose the hypothesis of a causal link between glyphosate exposure and risk of LHC. On balance, such analogies do not strengthen or weaken a conclusion of causality.

In summary, although none of the Bradford Hill viewpoints can establish or disprove causality, we did not find compelling evidence in support of causality based on any of the nine viewpoints. Thus, on balance, the existing epidemiologic evidence does not favor a causal effect of glyphosate on NHL, HL, MM, leukemia, or any subtype of these malignancies.

Discussion

Our meta-analysis yielded borderline significant RRs of 1.3 and 1.4 between glyphosate use and risk of NHL and MM, respectively, and no significant association with risk of HL or leukemia. Based on more fully adjusted RRs, our NHL meta-RR of 1.3 (95% CI = 1.0-1.6) was weaker than that reported by Schinasi and Leon^[11] (RR = 1.5, 95% CI = 1.1-2.0). The largest meta-RR of 2.5 (for hairy-cell leukemia) and the only meta-RR with a lower 95% confidence limit that excluded 1.0 (for B-cell lymphoma) were based on only two studies each, and the maximum number of studies contributing to any meta-analysis was six. The few studies with available data did not consistently detect positive exposure-response trends between quantitative measures of glyphosate use and risk of any LHC.

Consideration of the available epidemiologic evidence in light of the Bradford Hill viewpoints does not substantiate a causal relationship between glyphosate exposure and risk of any type of LHC. A conclusion in favor of causality also is undermined by the studies' methodological limitations, which could reasonably account for at least part of the observed associations. These limitations include exposure misclassification (which may differ by outcome status especially in case-control studies, which constitute nearly all available studies), selection bias (due to differential enrollment, follow-up, or data completeness), poor adjustment for confounding (by other agricultural exposures, for instance), small numbers (which lead to low statistical power as well as a higher probability that a statistically significant finding is false^[97]), and potential reporting and publication bias. Although underpowered statistical tests did not formally detect publication bias, we identified several examples of studies with available data that did not report associations between glyphosate use and LHC risk, and these unreported associations were most likely null.

Underpowered statistical tests also generally did not detect heterogeneity of results among studies, except for chronic lymphocytic leukemia/small lymphocytic lymphoma and MM. Nevertheless, our sensitivity analysis revealed some evidence of stronger associations with NHL risk in studies based in Sweden and those that ascertained cases in the 1980s, whereas the meta-RRs for studies that ascertained cases in the 2000s were



close to the null and statistically non-significant The stronger association with NHL diagnosed in the 1980s raises questions about whether glyphosate, an agent first introduced in 1974 in the United States and Europe, could plausibly cause lymphoma less than a decade later. However, deliberation on the potential induction time requires an understanding of the presumed mechanism of carcinogenesis, which is unknown for glyphosate. The classification system for lymphoid tumors underwent major changes in 1994 and 2001, [20] such that the definition of NHL as a disease entity is not entirely comparable between recent studies and those conducted in the 1980s. Study quality also may have improved over time, for example, due to refinements in survey design, interviewing techniques, data management, and other methods to augment data integrity.

The stronger association in Swedish studies probably is not explained by geographical differences in glyphosate use or effect modifiers related to NHL risk. One possible explanation is that of the six NHL studies, only the two Swedish studies^[14,15] compared subjects who used glyphosate with those who did not use any pesticides as the reference group, whereas the other studies defined the reference group as those who did not use glyphosate in particular. Comparisons with subjects who do not use any pesticides are more likely to be confounded by other pesticides and agricultural exposures.

Meta-analysis can be problematic when applied to observational epidemiology. [21,22] Meta-analysis increases statistical precision by combining results from studies that may differ substantially in terms of source population, exposure and outcome assessment and classification, control for confounding, and other key characteristics. In the presence of such heterogeneity, even if not detectable using formal statistical tests, a single summary estimate may not be scientifically meaningful. Additionally, even when studies are statistically homogeneous, meta-analysis may not yield valid results, since this technique cannot overcome problems in the design and conduct of the underlying studies. Instead, given that bias can seldom be ruled out and unmeasured and uncontrolled confounding can never be eliminated from observational epidemiologic studies, modest meta-RRs detected across multiple studies may simply be due to shared biases, rather than a true association. [21] As stated earlier, the purpose of meta-analysis is not to evaluate whether associations are causal. We conducted a meta-analysis primarily for comparison with published findings.

Considering the shortcomings of the existing literature, what can be done to shed further light on whether glyphosate causes LHC in humans? Perhaps the foremost need is better exposure assessment. Self-reported information on use of specific pesticides, unless validated by comparison with sales records (which most likely would need to be collected prospectively, and might not be closely correlated with pesticide use) or other objective documentation, is not sufficiently accurate and reliable to yield credible estimates of association, especially exposure-response trends. Urinary glyphosate levels would provide more accurate and quantitatively detailed information on biological dose of glyphosate received, but would probably have to be measured repeatedly to reflect long-term exposure.

Information about temporal aspects of glyphosate exposure, such as the putative induction time since first use of

glyphosate, duration of use, and time since last use, could help to shed light on the exposure-outcome relationship. Results from additional prospective cohort studies are necessary to alleviate concerns about selection and reporting bias in case-control studies.

More specific outcome classification also is needed. Only two studies [14,17] examined associations between glyphosate use and more than one histological subtype of NHL, despite growing evidence of important etiologic heterogeneity among NHL subtypes. [74] Information on NHL subtypes also is available in the Agricultural Health Study, [66] and publication of risk associations with glyphosate is anticipated. Risk factors for HL and leukemia also are known to differ by subtype, [76,77] yet no studies estimated associations with glyphosate separately for subtypes of these tumors. (Chronic lymphocytic leukemia and hairy-cell leukemia, which were analyzed as distinct outcomes, are classified as NHL subtypes. [20]) Large, probably pooled studies with histopathological data can determine whether associations with specific tumor subtypes might be obscured by analyzing overall NHL, HL, MM, or leukemia as a single disease entity.

Conclusion

In conclusion, we found marginally significant positive meta-RRs for the association between glyphosate use and risk of NHL and MM, and statistically null associations with HL and leukemia. A statistically significant positive meta-RR for B-cell lymphoma, but not other NHL subtypes, was calculated based on only two studies. Combining these results with recognition of the methodological weaknesses of the small number of existing studies and an overall body of literature that is not strong, consistent, temporally unambiguous, or indicative of a positive biological gradient, we determined that no causal relationship has been established between glyphosate exposure and risk of NHL, HL, MM, leukemia, or any subtype of LHC.

Acknowledgments

The authors thank John Acquavella and Thomas Sorahan for their thoughtful comments on earlier drafts of this manuscript, and Bernard Beckerman for his technical review of the tables.

Funding

This work was supported by Monsanto Company, the original producer and marketer of glyphosate formulations.

Disclosure statement

The sponsors were provided the opportunity to review the manuscript prior to journal submission, but inclusion of their suggestions was left to the discretion of the authors, who retained sole control of the manuscript content and the findings. Statements in this paper are those of the authors and not those of the authors' employer or the sponsors. The authors are employed by Exponent, a scientific research and consulting firm that provides services for private and governmental clients, including on projects concerning glyphosate and other pesticides. In the past five years, Ellen Chang has provided consulting services through Exponent on behalf of Monsanto Company on other issues, and she also has provided consulting services on other pesticides and lymphohematopoietic cancers for other clients.



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Appendix

Literature search methods

The authors conducted a search of MEDLINE via PubMed using the following search string, which includes Chemical Abstracts Service (CAS) Registry Numbers for glyphosate and its salts:

(glyphosat* OR glifosat* OR glyfosat* OR gliphosat* OR Roundup OR Round-up OR 1071-83-6 OR 38641-94-0 OR 70901-12-1 OR 39600-42-5 OR 69200-57-3 OR 34494-04-7 OR 114370-14-8 OR 40465-66-5 OR 69254-40-6 OR (aminomethyl w phosphonic*) OR 1066-51-9 OR pesticid* OR herbicid* OR organophosphorus compounds [MeSH] OR pesticides [MeSH] OR herbicides [MeSH]) AND (leukemi* OR leukaemi* OR lymphoma* OR NHL OR lymphopoietic OR hemato* OR hematopoie* or hematolog* OR lymphoid OR myeloid OR myeloma OR leukemia [MeSH] OR lymphoma [MeSH] OR multiple myeloma [MeSH]) AND (cases OR controls OR case-control OR cohort).

As of June 23, 2015, this search string identified a total of 11,755 articles in PubMed. We conducted additional targeted searches in PubMed, Web of Science, and Google Scholar using simpler keyword combinations such as (glyphosate AND lymphoma), (pesticides AND lymphoma), and (herbicides AND lymphoma). References also were identified from the bibliographies of recent review articles.

Altogether, a total of 12,709 articles were identified from these combined sources (Fig. A1). Based on a review of titles and abstracts, 321 articles were identified as potentially containing estimates of the association between glyphosate exposure and LHC risk, and were obtained for further evaluation. Forty-seven of these articles contained the word "glyphosate" or "Roundup" (or alternative spellings of these terms) in the text; as specified earlier, articles that did not mention glyphosate were ineligible for inclusion. Following a review of the full text of each of the 47 articles mentioning glyphosate, 19 articles were ultimately deemed eligible for inclusion.

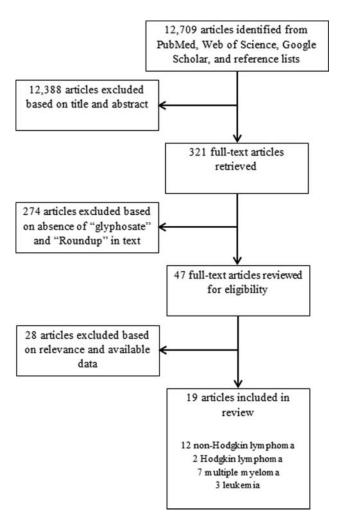


Figure A1. Flow chart of literature identification and selection process.